

18 June 2019

**Shellharbour City Council**  
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Attention: **Joel Coulton**  
Waste and Recovery Manager

**May 2019 Quarterly Environmental Monitoring – Dunmore Recycling and Waste Disposal Depot, Dunmore, New South Wales.**

## Introduction

Environmental monitoring is undertaken on a quarterly basis at the Dunmore Recycling Waste Disposal Depot, Dunmore, NSW (the site), in accordance with Environment Protection Licence (EPL) No. 5984. The monitoring includes sampling groundwater bores, a leachate pond, surface water bodies, a dust gauge and landfill gas at the landfill surface to detect any potential impacts of land filling activities on the environment.

## 1 SCOPE OF WORKS

On 14 May 2019, groundwater, surface water, leachate, gas and dust samples were collected in and around the site.

Groundwater samples were collected from nine monitoring bores (BH1c, BH2, BH3, BH4, BH13, BH14, BH16, BH20 and BH20s). At BH10 only the standing water level (SWL) was measured and no sample was taken. Surface water was collected from the leachate pond (LP1), three on site retention ponds (SWP1, SWP2, and SWP4) and Rocklow Creek at four points (SWC2, SWC\_Up, SWC\_Down and SWC\_Down\_2). Sampling was not taken at BH19 as the blockage encountered during the August monitoring round was still present. No sample was collected from BH15 as the access point was overgrown with trees and presented a potential WHS risk to field staff. No sample was taken at SWP5 as the retention pond was dry. Sampling locations are shown on **Figure 1 (Attachment 1)**.

A dust gauge bottle was collected to the north of the site (DDG) and a gas walkover of all site buildings and the landfill cap was also undertaken. Landfill gas was measured in the field using a Inspectra Laser Unit (ILU) and a GA5000 Landfill Gas Analyser (GA5000).



## 2 FIELD MEASUREMENTS

Prior to purging, monitoring bores were measured for SWL. During sampling, field measurements were taken including pH, electrolytic conductivity (EC), oxidation/reduction potential (ORP), dissolved oxygen and temperature. Colour and odour of water samples were also noted. Field measurements recorded for each location are presented in **Table 1 (Attachment 2)**.

All sampling was undertaken in accordance with Environmental Earth Sciences NSW (2011) *Soil, Gas and Groundwater sampling manual*.

## 3 LABORATORY ANALYSIS

The following analyses were undertaken for site groundwater and surface water during the May 2019 monitoring event:

- groundwater – ionic balance (pH, total dissolved salts (TDS), sodium, calcium, potassium, magnesium, fluoride, chloride, ammonium, sulfate, bicarbonate, phosphate and nitrate), total organic carbon (TOC), biological oxygen demand (BOD), total and soluble iron, and soluble manganese;
- surface water (SWC\_Up, SWC\_Down and SWC\_Down\_2) – ionic balance, total and soluble iron, turbidity, nitrate, ammonium and bicarbonate;
- surface water (SWC2) – ammonium, nitrate, bicarbonate and total and soluble iron;
- surface water SWP1, SWP2 and SWP4 – ionic balance, total and soluble iron and turbidity;
- additional analyses for SWP4 – TOC and BOD; and
- leachate tank (LP1) – ionic balance, TOC, BOD, total and soluble iron, soluble manganese, turbidity, faecal coliforms and E.Coli.

Water samples and the dust sample were sent to Sydney Analytical Laboratories (SAL) for inorganic chemical analyses and to Sonic Healthcare for faecal coliforms and Escherichia coli (E.Coli) counts. All laboratories are NATA accredited for the methods used.

The inorganic laboratory results for groundwater and surface water are shown in **Table 2 and Table 3 (Attachment 2)**. Calculated ratios of principal ions are presented in **Table 4 (Attachment 2)**.

## 4 RESULTS AND DISCUSSION

### 4.1 Groundwater flow

Inferred groundwater contours from the May 2019 standing water level (SWL) measurements are illustrated in **Figure 2 (Attachment 1)**. These were calculated using SWLs from surveyed bores. Groundwater flow direction was towards Rocklow Creek in a southerly direction similar to previous monitoring events.

Cumulative rainfall for March 2019 (157.2 mm), April 2019 (37.2 mm), May 2019 (7 mm) and June 2019 (56.2 mm) was 257.6 mm (BOM – Albion Park Wollongong Airport weather station) and slightly above 1999-2019 mean rainfall for this period of the year. Groundwater levels decreased in seven of the nine monitoring wells. The average of the measured standing water levels throughout the site have decreased by ~0.06 mAHD from 0.40 mAHD in February 2019 to ~0.46 mAHD.

### 4.2 Groundwater

#### 4.2.1 Groundwater sampling locations impacted by leachate

Field and laboratory results from the May 2019 sampling round, specifically from bores BH1c, BH2, BH3, BH4, BH20 and BH20s displayed chemistry that can be related to leachate impact with high levels of potassium, ammonium and nitrate. Leachate interaction is demonstrated by elevated concentrations of non-native potassium ( $K^+$ ), ammonium ( $NH_4^+-N$ ) and nitrate ( $NO_3^-$ ) relative to native sodium ( $Na^+$ ), calcium ( $Ca^{2+}$ ) and magnesium ( $Mg^{2+}$ ). This comparison is known as the leachate to non-leachate (L/N) ratio.

An L/N ratio >10 may be indicative of leachate impact depending on the combination with other indicators such as odour, colour, BOD and bicarbonate whereas a significant impact is likely to correspond with a ratio of >20 (**Table 4, Attachment 2**).

Bore BH1c is located near the old unlined landfill cell and intercepts leachate within the cell. As such the chemical signature of this bore has historically contained elevated leachate indicators in comparison to other monitoring bores (**Schoeller plot BH1 a/b/c, Attachment 3**). This continued during the current monitoring event and the groundwater was found to have a yellow brown colour and strong ammonia odour noted in combination with elevated TDS (4010 mg/L),  $K^+$  (220 mg/L) [resulting in low Ca/K – 1.15] and  $NH_4^+-N$  (325 mg/L) concentrations. The very low levels of dissolved oxygen (0.20 ppm, **Table 1**) and presence of soluble  $Fe^{2+}$  (3.5 mg/L) indicate an anaerobic environment and biochemical demand in response to microbial respiration. BOD has fluctuated since the bore was installed, ranging from 850 mg/L to 6 mg/L. BODs have remained at similar levels (6-25 mg/L) since August 2016 with a trend of small decreases up to the May 2019 BOD concentration of 12 mg/L. Further evidence of microbial activity and respiration of nitrogen species in groundwater is elevated  $HCO_3^-$  resulting in a low Cl/ $HCO_3^-$  ratio of 0.48 (**Table 4**). This suggests some degradation of the leachate plume, and the organic nitrogen species therein, has occurred in this monitoring bore.

Bore BH2 is located down gradient from the old unlined landfill cell. Historically elevated levels of  $NH_4^+-N$  indicate some leachate impact at this location.  $NH_4^+-N$  concentration at

BH2 showed an increasing trend since 2010 and reached its historical maximum in August 2017 and November 2017 (49 mg/L in both months).  $\text{NH}_4^+$ -N concentration increased slightly in the May 2019 monitoring round to a level of 41 mg/L (up from 39 mg/L in November 2018). Bicarbonate ( $\text{HCO}_3^-$ ),  $\text{Na}^+$  and  $\text{Mg}^{2+}$  concentrations in groundwater have shown an increasing trend since May 2008 (**Table 2** and **Schoeller plot BH2, Attachment 3**). Calcium ( $\text{Ca}^{2+}$ ) and potassium ( $\text{K}^+$ ) concentrations slightly decreased since the last monitoring round (**Table 2** and **Schoeller plot BH2, Attachment 3**). These two ions (and chlorine ( $\text{Cl}^-$ )) have generally followed the same trend observed in May 2019 monitoring round since September 2015. Low oxygen and negative redox (**Table 2**) continue to suggest microbial respiration and therefore degradation of the leachate is occurring at this location. Additionally, a slight sweet odour was noted.

Groundwater from bore BH3 reported an increase in concentration of native ions ( $\text{Na}^+$ ,  $\text{Mg}^{2+}$ ,  $\text{K}^+$  and  $\text{Ca}^{2+}$ ). Non-native nitrogen species increased in May 2019 with  $\text{NH}_4^+$ -N concentration rising to 27 mg/L from 19 mg/L in February 2019.  $\text{NO}_3^-$  recorded an increase in concentration to 105 mg/L from 56 mg/L in the previous monitoring event (February 2019). The concentration of  $\text{HCO}_3^-$  also increased from 290 mg/L to 490 mg/L. The L/N ratio (60.29 %) had increased since the February 2019 monitoring round (51.94 %). It was reported and verified during the drilling of BH3 that old unconfined waste dumps were in the vicinity of bore BH3 (outside the designated cells near bore BH2). In the previous monitoring report, it was suggested that elevated nitrogen species concentrations and a proportional increase of L/N ratio was due to the relatively high rainfall recorded in October and November 2018 and subsequent leaching of nitrogen species from the overlying unconfined waste in the vicinity of BH3 through the soil profile and into groundwater. Elevated L/N ratios after significant rainfall has been observed over the historical data range. The best example of this trend was observed during the August 2013 monitoring round, which was undertaken following a cumulative rainfall of 390.2 mm for May, June and July comparative to the 198.5 mm mean rainfall for 1999-2018 for that period. BH3's historic peak L/N ratio (208.90%) was recorded that monitoring round.

Bore BH20 is located down gradient of the landfill, leachate ponds and shallow old landfill. This bore was positioned to assess the chemical characteristics on the boundary of the landfill site. Field observations at bore BH20 recorded a negative redox (-89 mV) with clear groundwater and no odour. The L/N ratio (17.54 %) in the May 2019 round had decreased from the February 2019 value (22.12 %) but was still considered significantly elevated. The TDS remained relatively low (rising from 815 mg/L to 855 mg/L) making the L/N susceptible to natural variations or fluctuations in chemistry. Chemical characteristics of the bore show groundwater is low in  $\text{Na}^+$ , with a moderate Ca/K and K/TDS ratio (**Table 4**). Ammonium levels remained elevated at 21 mg/L however other landfill indicators were low or absent.

Bore BH20s is located directly adjacent to BH20 but at a shallower depth. Screened intervals of BH20 and BH20s are 6.0-9.0 mBGL and 1.5-4.5 mBGL respectively. Similarly, this bore was positioned to compare the chemical characteristics on the boundary of the landfill site in order to locate potential transport pathways to Rocklow Creek. In May 2019, field measurements at bore BH20s recorded a positive redox (24 mV), indicative of an oxidative atmosphere. Groundwater was clear, and no odour was detected.  $\text{NO}_3^-$  concentrations increased from 33 mg/L in February 2019 to 54 mg/L in May 2019. The increase in  $\text{NO}_3^-$  led to a slight decrease in L/N ratio (from 70.46 % to 69.32 %), however this value was still elevated and is indicative of potentially high leachate impact at this site. TDS is relatively low (810 mg/L) making the L/N susceptible to natural variations or fluctuations in

chemistry. Chemical characteristics of the bore show groundwater was low in  $\text{Na}^+$ , with a moderate Ca/K and K/TDS ratio (**Table 4**). As observed within BH3, the relatively high rainfall from March to June 2019 may have impacted the nitrogen species within BH20s, causing leaching of nitrogen species from the soil into the groundwater, resulting in elevated  $\text{NO}_3^-$  concentrations. Ammonium levels (1.1 mg/L) have increased significantly since February 2019 (0.1 mg/L) and remain lower than those seen at the deeper BH20 bore. It was previously thought that high nitrate levels in this shallower bore location was indicative of nitrification throughout the soil profile, however, continued monitoring at this location will be necessary to determine potential leachate transport pathways to Rocklow Creek.

#### 4.2.2 Remaining groundwater sampling locations

During the May 2019 monitoring round, ionic chemistry indicated that bores BH4, BH13, BH14 and BH16 only displayed slight to no leachate influence. Chemical composition of each of these bores has been depicted in **Schoeller plots** in **Attachment 3**.

The L/N ratio at bore BH4 showed a slight decrease in the May 2019 round (8.89 %) — in November 2018 monitoring round the L/N ratio was 9.59 %. The L/N ratio at this location had not previously exceeded 10% since May 2003.  $\text{NH}_4^+$ -N levels increased from 6.70 mg/L to 8.9 mg/L, however there was a sharp decrease in nitrite ( $\text{NO}_2^-$ ) from 9.6 mg/L to 0.23 mg/L, indicative of a decrease in the nitrification process and transformation of  $\text{NH}_4^+$ -N to  $\text{NO}_2^-$ . BH4 is placed on the border of an historic shallow landfill site and down gradient of landfilling activities. This area should be continually monitored to determine water quality in this area.

Bore BH13 is in close proximity to a former night soil area (**Figure 1**). A slight residual leachate influence has been apparent at this location in the past. Analysis of chemical data from the May 2019 monitoring round shows an increase of L/N ratio of 12.03 % from 10.49 % in February 2019. Concentrations of  $\text{NO}_3^-$  continued to decrease from 3.10 mg/L to 0.35 mg/L. Large fluctuations in  $\text{NO}_3^-$  have previously been observed in the historic data, however, chemical composition of the groundwater has generally remained consistent since monitoring began in 2002 (**Schoeller plot BH13, Attachment 3**).

The L/N ratio at bore BH14 remained stable in the May 2019 round (5.14 %) — in February 2019 monitoring round the L/N ratio was 5.47 % at this location.  $\text{NO}_3^-$  concentration remained stable at low levels however  $\text{NO}_3^-$  levels have been historically high at this location.  $\text{NH}_4^+$ -N concentrations increased slightly from 1.70 to 1.80 mg/L however remain below the site criteria of 1.88 mg/L. Bore BH14 is strategically placed down gradient of landfilling activities and should be continually monitored to determine the water quality in this area given its history of leachate impact.

Bore BH16 is located in a swampy area with groundwater field observations recording a murky colour and a minor leachate odour. The sampled redox potential indicates a reducing environment (-161 mV), which may have an influence on the historical dominance of  $\text{NH}_4^+$ -N over  $\text{NO}_3^-$ . This round  $\text{NH}_4^+$ -N concentrations increased slightly to 0.2 mg/L from 0.1 mg/L. Groundwater sampling in May 2019 indicated limited to no leachate impact at BH16 despite a slightly elevated L/N ratio of 12.81 %. Bore BH16 is located close to a drainage channel where offsite impacts can readily influence the chemical characteristics of the shallow groundwater and should continue to be monitored for fluctuations.

### 4.2.3 Groundwater site criteria exceedances

NH<sub>4</sub><sup>+</sup>-N concentrations above threshold levels (1.88 mg/L) (ANZECC, 2000) were reported in groundwater from bores BH1c (325 mg/L), BH2 (41 mg/L), BH3 (27 mg/L), BH4 (8.9 mg/L), BH13 (2.8 mg/L) and BH14 (14 mg/L). Nitrate (NO<sub>3</sub><sup>-</sup>) was reported above guideline thresholds (10.6 mg/L) (ANZECC, 2000) at BH3 (105 mg/L).

## 4.3 Surface water monitoring

During the May 2019 monitoring round, samples from Rocklow Creek (SWC2, SWC\_Up, SWC\_Down and SWC\_Down\_2) and three surface water ponds (SWP1, SWP2, and SWP4) were collected. Results of surface water analysis (**Table 2** and **Table 3**) indicate that concentrations of ions were within the historical ranges. As surface water ponds are intended to retain any surface water migrating towards Rocklow Creek, the detection of chemical constituents that may be associated with landfill leachate are expected.

NH<sub>4</sub><sup>+</sup>-N levels detected at SWP1 (0.30 mg/L) increased since the previous monitoring event. Ongoing minor leachate impact has been indicated by consistent L/N ratios > 10% and < 20%. Elevated concentrations of soluble iron and a negative redox potential are indicative of a reducing environment which may have contributed to historical low levels of dissolved oxygen and the production of NH<sub>4</sub><sup>+</sup>-N.

Surface water sampled at SWP2 showed little to no leachate impact. The surface water pond collects runoff from around the site and potential impacts from site activities are often observed. NH<sub>4</sub><sup>+</sup>-N concentration remained low at 0.10 mg/L however, fluctuating NH<sub>4</sub><sup>+</sup>-N is common at this location with previous monitoring events ranging between 0.01 and 15 mg/L. NO<sub>3</sub><sup>-</sup> concentrations were below the limit of detection (0.10 mg/L). All chemical parameters at this location are within historical ranges.

NH<sub>4</sub><sup>+</sup>-N concentration at SWP4 increased significantly from <0.10 mg/L to 3.20 mg/L and exceeded the trigger level of 1.88 mg/L. NO<sub>3</sub><sup>-</sup> levels also increased above the limit of detection to 0.80 mg/L in May 2019. The increase in NH<sub>4</sub><sup>+</sup>-N and NO<sub>3</sub><sup>-</sup> concentrations is indicative of a slow down in the natural process of nitrification by which NH<sub>4</sub><sup>+</sup>-N naturally attenuates. All chemical parameters at this location are within historical ranges.

The four surface water creek sites SWC2, SWC\_Up, SWC\_Down and SWC\_Down\_2 (**Figure 2**) were also sampled during the May 2019 sampling event. SWC\_Down\_2 is still a relatively new sampling location which was established in order to detect potential leachate impacts to Rocklow Creek originated from the eastern portion of the site (Environmental Earth Sciences NSW, 2017). SWC\_Up, SWC\_Down and SWC\_Down\_2 had high concentrations of TDS (23,600 mg/L), notably Na<sup>+</sup> and Cl<sup>-</sup> (**Table 2**); this is due to the tidal nature of these waters and differentiates them from landfill groundwater / surface water.

The low nutrient and L/N ratios within Rocklow Creek indicated that there was no leachate impact within Rocklow Creek. All surface water creek sampling sites (SWC2, SWC\_Up, SWC\_Down and SWC\_Down\_2) had concentrations of NH<sub>4</sub><sup>+</sup>-N and NO<sub>3</sub><sup>-</sup> below the ANZECC (2000) guidelines, with the exception of NH<sub>4</sub><sup>+</sup>-N at SWC2 (3.20 mg/L) above the threshold level of 1.88 mg/L. A review of the NH<sub>4</sub><sup>+</sup>-N concentrations at SWC2 since 1998, shows the NH<sub>4</sub><sup>+</sup>-N concentration has fluctuated ranging from 0.1 to 4.9 mg/L. This is the fifth time, in 21 years, NH<sub>4</sub><sup>+</sup>-N concentration has exceeded the threshold level of 1.88 mg/L, with the highest

result of 4.5 mg/L reported in November 2011. All four sites will continue to be monitored to ensure leachate is not impacting upon the Rocklow Creek.

#### 4.4 Monitoring of Leachate Tanks

The chemistry of leachate water at the Dunmore Recycling and Waste Disposal Depot is significantly different when compared to the surface and groundwater chemistry of non-leachate influenced bores. This is demonstrated through comparison of chemical data presented in **Table 2**. In particular TDS, BOD, TOC, Na<sup>+</sup>, NH<sub>4</sub><sup>+</sup>-N, K<sup>+</sup>, HCO<sub>3</sub><sup>-</sup>, PO<sub>4</sub><sup>3-</sup> and total iron concentrations are generally elevated in leachate pond water compared to other monitoring bores (**Schoeller Plot LP1, Attachment 3**). Ionic ratios (**Table 4**) such as low Ca/K (0.60) and high Na/Ca (10.30) and L/N (83.51) ratios represent landfill leachate chemical characteristics. These chemical characteristics have been relatively stable over the past 10 years of monitoring.

Laboratory analysis did not detect faecal coliforms and E.Coli during this round, a significant decrease on the concentrations of 170 CFU/100 ml and 140 CFU/100mL respectively reported in February 2019 (**Table 3**). Fluctuation in these concentrations in leachate tanks are common and thus dermal contact with these waters should continue to be avoided due to historic levels of elevated concentrations and the corresponding health concerns relating to high microbial counts.

#### 4.5 Quality assurance/quality control

For quality assurance and quality control the following precision and reliability measures were calculated. The charge balance difference between the summed total of anions against cations (milli-equivalent units) was in the range of 0.52 % to 3.57 %. The results are a good indication that all major cations and anions present in the groundwater have been analysed and accounted for, providing confidence in the laboratory results obtained.

Field and laboratory practices were further evaluated by comparing the difference between field and laboratory pH and field measured electrical conductivity (EC) against laboratory total dissolved salts (TDS). The range of most relative percent difference (RPD) of field to laboratory pH measurements was between 0.14 % and 9.02 %. The relationship between the field determined EC and laboratory measured TDS relationship ranged between 0.50 and 0.77.

RPDs between the intra-laboratory duplicate and the primary sample taken at bore BH4 were all within the acceptable RPD criteria. Thus, the data is considered reliable (**Table 6**).

#### 4.6 Gas monitoring

Landfill gas was measured in the field using an Inspectra Laser Unit (ILU) and a GA5000 Landfill Gas Monitor (GA5000). Measurements were taken within and around all buildings in a 250 m radius from the current landfill cell as well as across the landfill cap (gas walkover grids of the May 2019 round are presented in **Figure 3**).

All readings were below the site-specific criteria outlined in EPL no. 5984 as the NSW EPA (2013) reporting threshold of 1.00 % v/v CH<sub>4</sub> within onsite buildings and therefore pose no direct risk. All readings were below the threshold concentration for closer investigation and

potential action (500 ppm or 0.05 % v/v, NSW EPA [2013], **Table 5**). Continued monitoring with both the GA5000 and ILU will be undertaken at quarterly monitoring events.

#### 4.7 Dust

Dust deposition levels to the north of the site were 1.1 g/m<sup>2</sup>/month total solids, which is below the accepted level of 4 g/m<sup>2</sup>/month (Australian Standards AS3580.10.1 and AS2724.1). Dust deposition levels are within historical ranges and will continue to be monitored to assess the closest sensitive receptor, houses located to the north-west of site.

## 5 CONCLUSION AND RECOMMENDATIONS

Groundwater behaviour across the site since the commencement of quarterly monitoring in 1992 has been generally consistent. As the plume beneath the site is relatively stable, changes in leachate behaviour into the future are not expected to be significant. Changes to site conditions such as stockpile locations, new landfill cells, new retention ponds and other earth works could potentially impact leachate behaviour on site.

The May 2019 monitoring round found a general decrease in L/N ratios at leachate impacted groundwater sampling locations due to decreases in nitrogen species concentrations, particularly at BH2, BH20 and BH20s. BH3 and BH13 noticed slight increases in L/N ratios.

Assessment of monitoring bores closest to Rocklow Creek, BH20 and BH20s, has detected the presence of leachate indicators despite the Rocklow Creek samples (SWC-Up, SWC-Down and SWC\_Down\_2), showing no affect. Although the historical data sets of these new bore locations are relatively limited, it appears that on-site activities are not significantly impacting Rocklow Creek. Surface water monitoring indicated that on site activities have had limited impact on water quality at locations SWP1, SWP2, SWP4 and SWP5. Assessment of Rocklow Creek sampling locations (SWC2, SWC-Up, SWC-Down and SWC\_Down\_2) reported no concentrations of NH<sub>4</sub><sup>+</sup>-N and NO<sub>3</sub><sup>-</sup> above the ANZECC (2000) trigger value.

Gas concentrations detected at all buildings assessed on site were below guidelines and therefore no action was required. It is recommended that monitoring continue with a FID or Inspectra Laser Unit and GA5000 Landfill Gas Monitor.

Depositional dust monitoring results continued to be below guidelines (Australian Standards AS3580.10.1 and AS2724.1) and will continue to be monitored to assess the impact that dust poses on nearby residential areas.

## 6 LIMITATIONS

This report has been prepared by Environmental Earth Sciences NSW ACN 109 404 006 in response to and subject to the following limitations:

1. The specific instructions received from Shellharbour City Council;
2. The specific scope of works is set out in PO117559 issued by Environmental Earth Sciences NSW for and on behalf of Shellharbour City Council.
3. May not be relied upon by any third party not named in this report for any purpose except with the prior written consent of Environmental Earth Sciences NSW (which consent may or may not be given at the discretion of Environmental Earth Sciences NSW);
4. This report comprises the formal report, documentation sections, tables, figures and appendices as referred to in the index to this report and must not be released to any third party or copied in part without all the material included in this report for any reason;
5. The report only relates to the site referred to in the scope of works being located at Dunmore Recycling and Waste Disposal Depot located at Buckleys Rd Dunmore, NSW;
6. The report relates to the site as at the date of the report as conditions may change thereafter due to natural processes and/or site activities;
7. No warranty or guarantee is made in regard to any other use than as specified in the scope of works and only applies to the depth tested and reported in this report;;
8. This report is not a geotechnical or planning report suitable for planning or zoning purposes; and
9. Our General Limitations set out at the back of the body of this report.

Should you have any queries, please do not hesitate to contact us on (02) 9922 1777.

For and on behalf of  
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## 7 REFERENCES

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# ENVIRONMENTAL EARTH SCIENCES GENERAL LIMITATIONS

## **Scope of services**

The work presented in this report is Environmental Earth Sciences response to the specific scope of works requested by, planned with and approved by the client. It cannot be relied on by any other third party for any purpose except with our prior written consent. Client may distribute this report to other parties and in doing so warrants that the report is suitable for the purpose it was intended for. However, any party wishing to rely on this report should contact us to determine the suitability of this report for their specific purpose.

## **Data should not be separated from the report**

A report is provided inclusive of all documentation sections, limitations, tables, figures and appendices and should not be provided or copied in part without all supporting documentation for any reason, because misinterpretation may occur.

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Understanding an environmental study will reduce exposure to the risk of the presence of contaminated soil and or groundwater. However, contaminants may be present in areas that were not investigated, or may migrate to other areas. Analysis cannot cover every type of contaminant that could possibly be present. When combined with field observations, field measurements and professional judgement, this approach increases the probability of identifying contaminated soil and or groundwater. Under no circumstances can it be considered that these findings represent the actual condition of the site at all points.

Environmental studies identify actual sub-surface conditions only at those points where samples are taken, when they are taken. Actual conditions between sampling locations differ from those inferred because no professional, no matter how qualified, and no sub-surface exploration program, no matter how comprehensive, can reveal what is hidden below the ground surface. The actual interface between materials may be far more gradual or abrupt than an assessment indicates. Actual conditions in areas not sampled may differ from that predicted. Nothing can be done to prevent the unanticipated. However, steps can be taken to help minimize the impact. For this reason, site owners should retain our services.

## **Problems with interpretation by others**

Advice and interpretation is provided on the basis that subsequent work will be undertaken by Environmental Earth Sciences NSW. This will identify variances, maintain consistency in how data is interpreted, conduct additional tests that may be necessary and recommend solutions to problems encountered on site. Other parties may misinterpret our work and we cannot be responsible for how the information in this report is used. If further data is collected or comes to light we reserve the right to alter their conclusions.

## **Obtain regulatory approval**

The investigation and remediation of contaminated sites is a field in which legislation and interpretation of legislation is changing rapidly. Our interpretation of the investigation findings should not be taken to be that of any other party. When approval from a statutory authority is required for a project, that approval should be directly sought by the client.

## **Limit of liability**

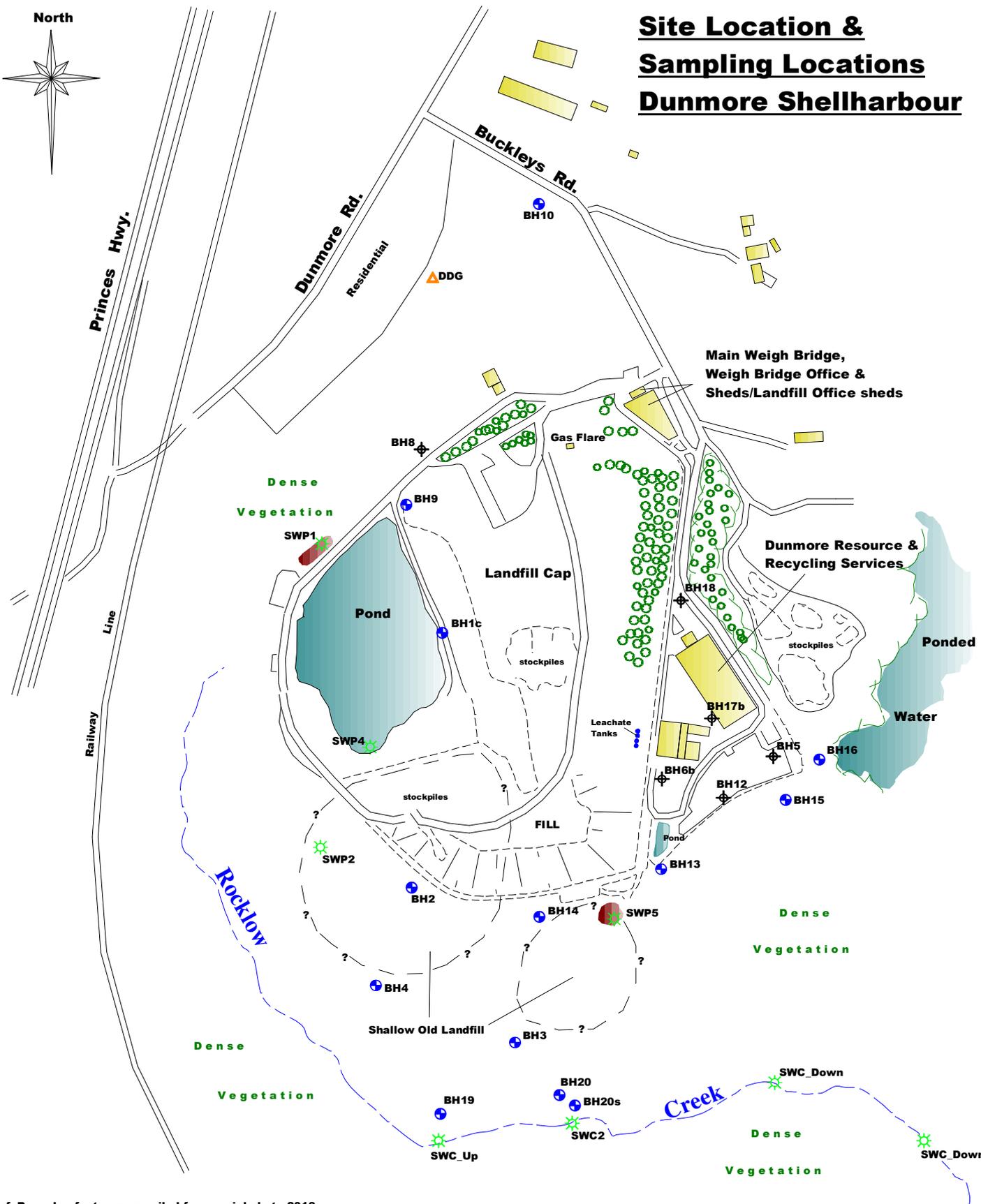
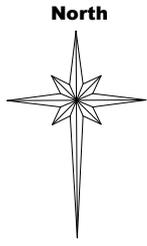
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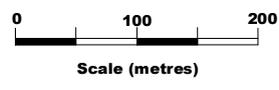
## APPENDIX A: FIGURES

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# Site Location & Sampling Locations Dunmore Shellharbour



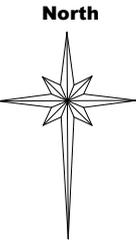
Ref. Baseplan features compiled from aerial photo 2018.



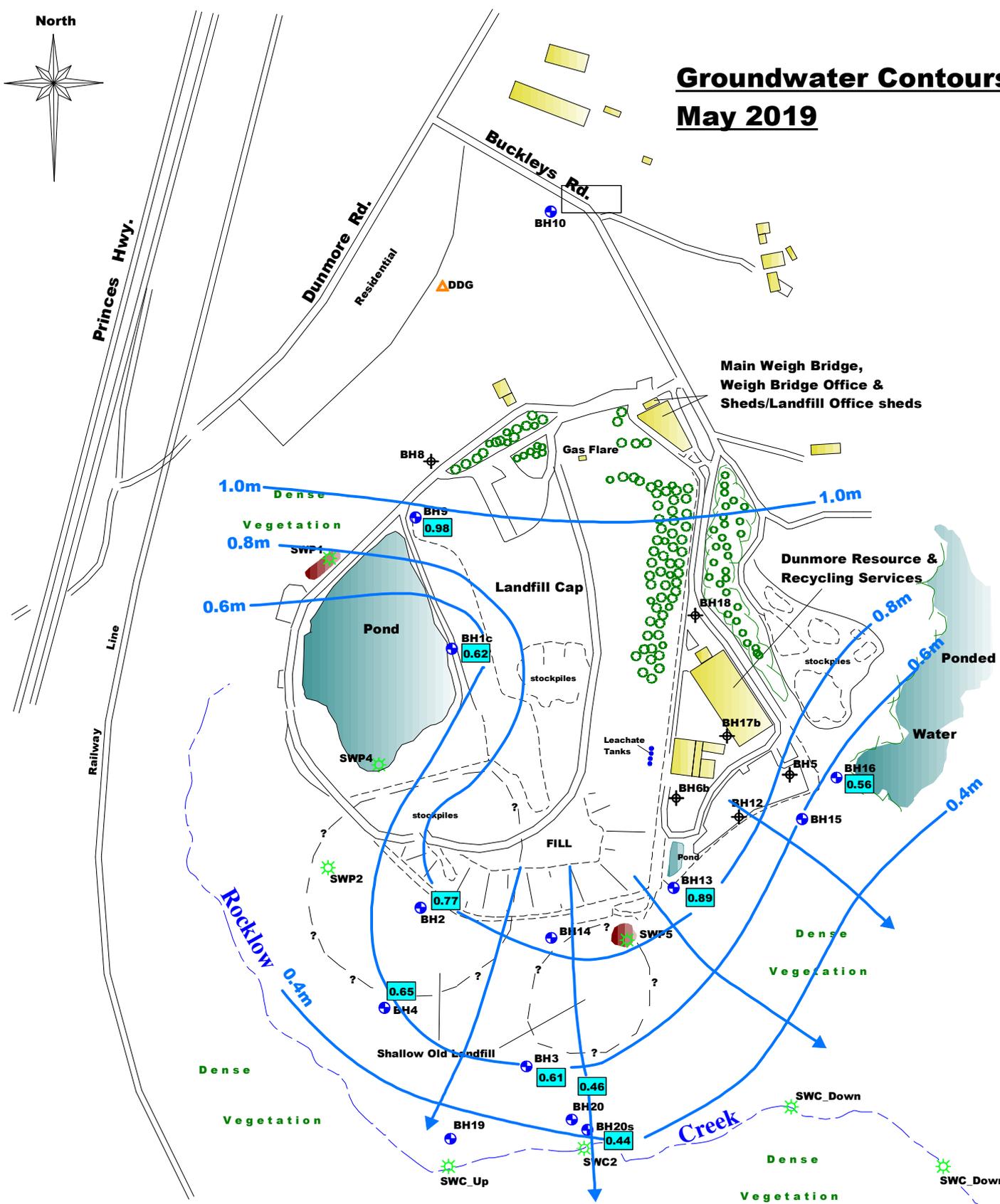
**Legend:**

- Bore locations
- ▲ Dust gauge locations
- ☼ Surface water locations
- ⊕ Decommissioned bores

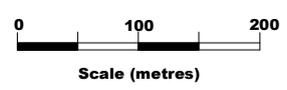
 <b>ENVIRONMENTAL EARTH SCIENCES</b> CONTAMINATION RESOLVED	<b>Title: Site Plan &amp; Sampling Locations</b>		
	<b>Location: Buckleys Road, Dunmore, Shellharbour, NSW</b>		
<b>Client: Shellharbour City Council</b>	<b>Job No. 118077</b>	<b>Date: May 2019</b>	<b>Figure 1</b>
	Drawn by: TRJ	Proj. Man. EG	



# Groundwater Contours May 2019



Ref. Baseplan features compiled from aerial photo 2018.

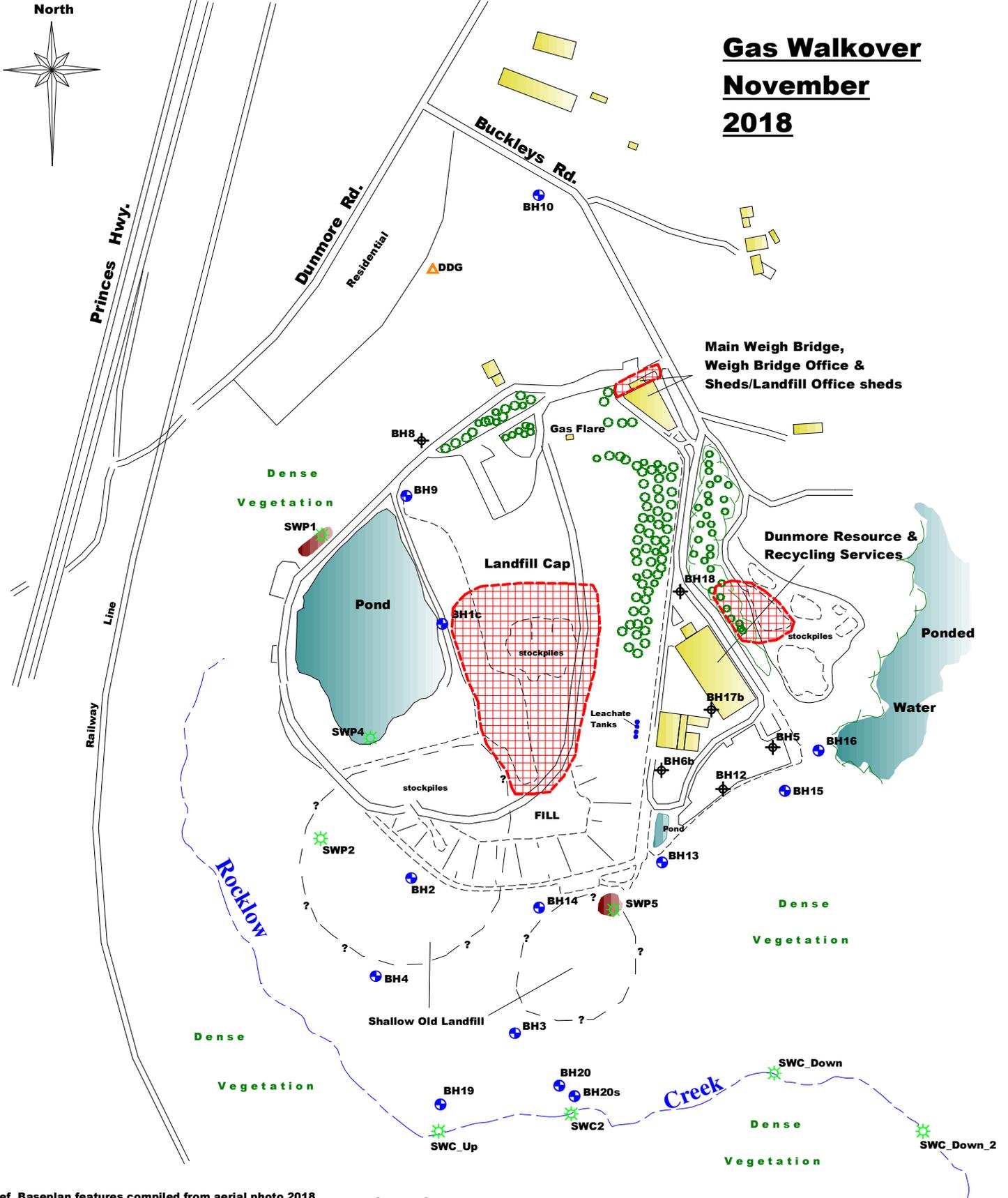


**Legend:**

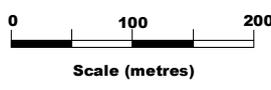
- Bore locations
- ▲ Dust gauge locations
- ★ Surface water locations
- ⊕ Decommissioned bores
- Inferred flow direction
- Inferred groundwater contour (m AHD)
- 0.46 SWL (standing water level m AHD)

 <b>ENVIRONMENTAL EARTH SCIENCES</b> CONTAMINATION RESOLVED	<b>Title:</b> Groundwater Contours - May 2019		
	<b>Location:</b> Buckleys Road, Dunmore, Shellharbour, NSW		
<b>Client:</b> Shellharbour City Council	<b>Job No. 118109</b>	<b>Date:</b> 13 June 2019	<b>Figure 2</b>
	<b>Drawn by:</b> TRJ	<b>Proj. Man.:</b> EG	

# Gas Walkover November 2018



Ref. Baseplan features compiled from aerial photo 2018.



**Legend:**

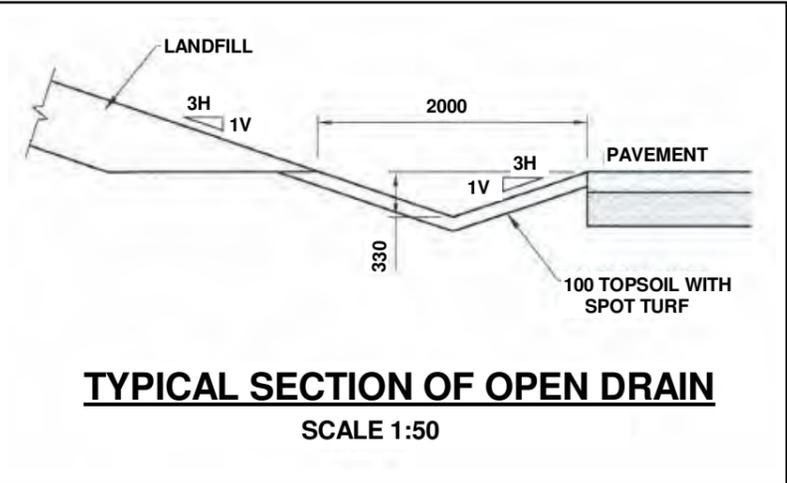
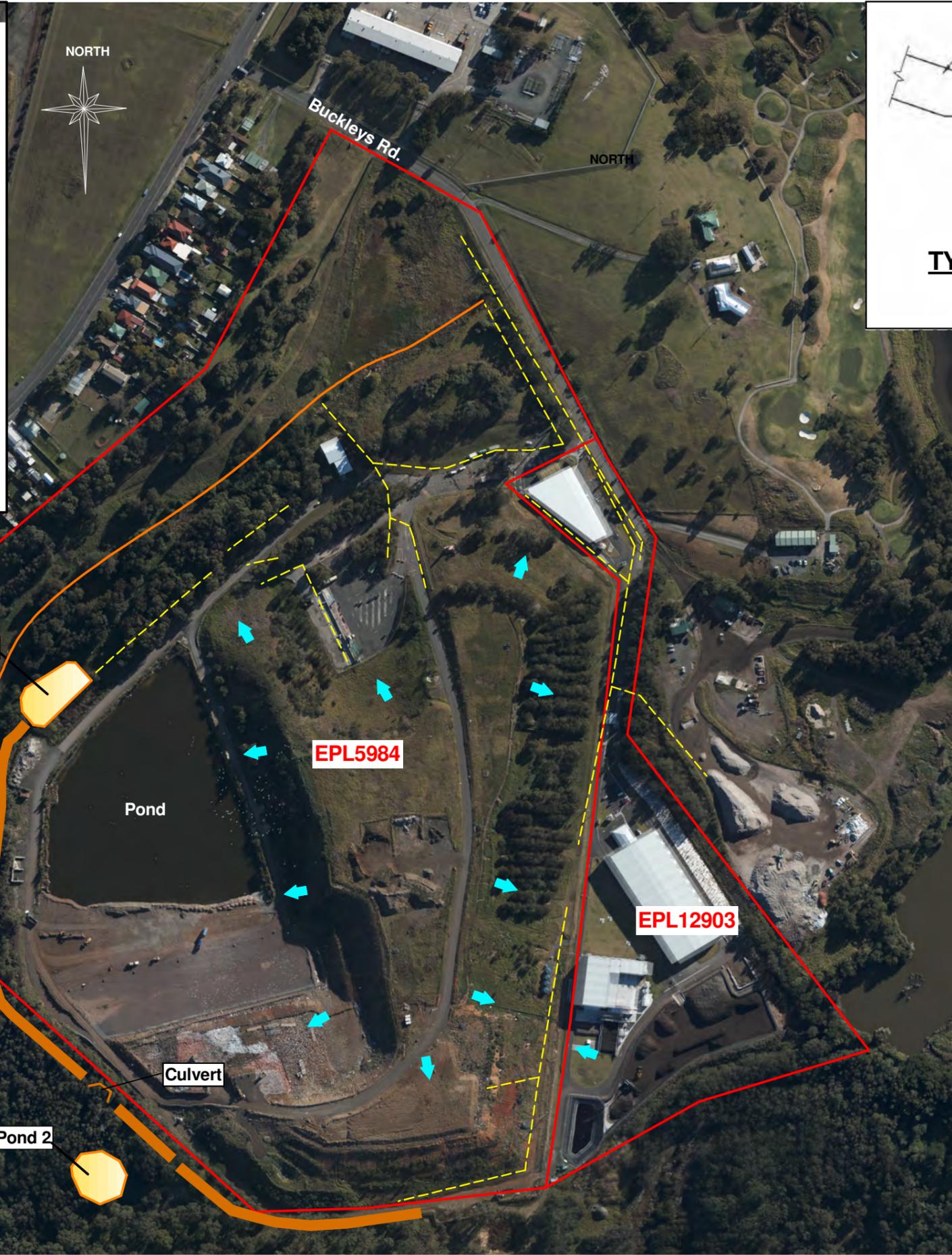
- Bore locations
- ▲ Dust gauge locations
- ★ Surface water locations
- ⊕ Decommissioned bores
- Gas walkover

 <b>ENVIRONMENTAL EARTH SCIENCES</b> CONTAMINATION RESOLVED	<b>Title:</b> Gas Walkover		
	<b>Location:</b> Buckleys Road, Dunmore, Shellharbour, NSW		
	<b>Client:</b> Shellharbour City Council	<b>Job No. 118077</b>	<b>Date:</b> May 2019
	<b>Drawn by:</b> TRJ	<b>Proj. Man.:</b> EG	<b>Scale:</b> As shown

**Figure 3**

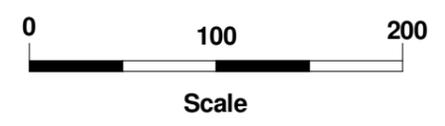


Regional Locality Map



Legend:

- licence boundary
- levee bank
- drainage layout
- approx. stormwater diversion drain
- ▶ - approx. stormwater flow direction



Ref. Based on updated aerial photograph 2018.

	Title: Surface Water Runoff	
	Location: Dunmore Recycling & Waste Disposal Depot, Shellharbour	
Client: Shellharbour City Council	Job number: 118109	
Drawn by: TRJ	Scale: As shown	Source: See Ref.
Proj Man: EG	Date: May 2019	Figure 4

## APPENDIX B: TABLES

---

**Table 1: Field measurements – May 2019**

Sample	SWL	SWL	pH	EC	ORP	Temp.	DO	Colour	Odour
Units	mAHD	Dip (m)	---	mS/cm	mV	°C	ppm	---	---
BH1c	0.624	3.33	6.73	7.33	-96	25.9	0.2	Yellow brown	Strong ammonia
BH2	0.767	4.025	6.89	3.096	-74.8	23	0.09	Yellow	Slight sweet
BH3	0.614	3.15	7.19	1.546	1.1	19.2	1.08	Clear	None
BH4	0.649	4.37	7.3	1.73	-70	19.6	0.04	Clear	Eggy odour
BH9	0.975	3.41	---	---	---	---	---	---	---
BH10	---	---	---	---	---	---	---	---	---
BH13	0.885	4.41	6.8	1.55	-29	21.5	1.3	Clear	None
BH14	0.855	4.86	6.8	2.04	14	22.1	1.2	Slight light brown	Slight sweet
BH16	0.56	0.82	7.29	0.362	-161	18.8	0.28	Murky	Egg odour
BH20	0.46	2.31	7.28	1.3	-89	19.2	0.74	Clear	None
BH20s	0.44	2.33	7.27	1.14	24	20.1	0.07	Clear	None
LP1	---	---	8.11	13.53	-32.4	18.3	4.06	Strong leachate	Black
SWC2	---	---	7.26	39.66	43.6	16.1	3.05	Clear	None
SWC_UP	---	---	7.11	38.16	-180	16.3	0.37	Clear	Hydrogen Sulphide
SWC_DOWN	---	---	6.8	41.63	0.93	16.5	2.05	Clear	None
SWC_DOWN_2	---	---	7.12	39.83	-39	16.1	3.34	Clear	None
SWP1	---	---	7.88	0.351	-20.1	13.3	0.16	Pond green	
SWP2	---	---	7.74	1.7	-68	13.4	1.47	Clear	None
SWP4	---	---	7.89	2.07	73	18	3.95	Murky green brown	
SWP5	---	DRY	---	---	---	---	---	---	---

**Notes:**

SWL Standing Water Level, measured to the top of the monument or casing; RL – reference level;

--- not measured;

N/A = Not applicable

DO = dissolved oxygen;

ORP = electron activity; and

EC= electrolytic conductivity

**Table 2: Water laboratory results – May 2019**

Sample	pH	TDS	Na	Ca	K	Mg	NH4-N	Cl	F	NO3	NO2	SO4	HCO3	PO4	TOC	BOD	Sol. Mn	Sol. Fe	Tot. Fe
		mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L
BH1c	7.1	4010	650	130	220	115	<b>325</b>	850	0.18	<0.1	---	15	3070	<0.1	175	12	0.14	<b>3.5</b>	19
BH2	7.1	1820	335	175	45	74	<b>41</b>	400	0.23	<0.1	---	120	1220	<0.1	65	2	0.50	<b>2.5</b>	12
BH3	7.2	775	70	180	35	27	<b>27</b>	200	<0.1	<b>105</b>	---	77	490	<0.1	13	<2	0.12	0.10	1.6
BH4	7.1	1100	145	190	24	36	<b>8.9</b>	215	<0.1	<0.1	0.23	155	675	<0.1	20	3	0.21	0.19	5.6
BH13	7.1	995	100	175	35	42	<b>2.8</b>	99	0.21	0.35	---	220	645	<0.1	23	<2	0.26	0.22	1.9
BH14	6.9	1290	205	195	20	57	1.8	235	0.39	1.7	---	85	920	<0.1	29	<2	0.40	0.16	4.5
BH16	7.2	215	33	16	7.9	15	0.2	44	0.32	<0.1	---	43	88	<0.1	16	4	0.07	0.26	6.0
BH20	7.4	855	49	180	32	34	<b>14</b>	120	0.12	0.13	---	240	440	0.21	20	<2	0.08	0.12	1.9
BH20s	7.4	810	46	120	87	39	1.1	54	0.10	54	---	190	425	<0.1	16	<2	0.07	0.07	0.12
LP1	8.5	7270	1300	110	360	93	<b>895</b>	1600	0.44	<0.1	---	105	5750	15	560	28	0.49	<b>3.7</b>	7.9
SWC2	---	---	---	---	---	---	<b>3.0</b>	---	---	0.22	0.36		210					0.13	1.5
SWP1	7.2	270	44	31	8.2	12	0.3	60	<0.1	0.49	---	11	175	0.12				<b>0.64</b>	1.0
SWP2	7.8	1260	280	81	27	50	0.1	330	0.13	1.0	---	170	545	<0.1				0.07	0.09
SWP4	7.7	1350	350	62	23	62	<b>3.2</b>	375	0.28	0.80	---	305	435	<0.1	33	3		0.09	0.35
SWC-UP	7.2	14900	4520	235	150	560	<0.1	8070	0.38	<0.1	---	1180	165	<0.1				0.16	16
SWC-DOWN	7.2	22500	6670	300	250	820	0.4	12300	0.39	0.35	---	1670	195	<0.1				0.09	0.24
SWC_DOWN_2	7.3	23600	6950	310	265	840	0.3	13000	0.39	0.31	---	1780	190	<0.1				0.05	0.32
ANZECC 2000	6.5-8.0	---	---	---	---	---	1.88*	---	---	10.6 <sup>#</sup>	---	---	---	---	---	---	---	0.3	---

Notes:

Results and guidelines are expressed in mg/L

SWC\_Do – SWC\_Down;

--- not analysed;

Guidelines levels from ANZECC (2000) – Australian and New Zealand guidelines for fresh and marine water quality for the protection of aquatic ecosystems;

\* - guideline from freshwater trigger values as total NH<sub>4</sub>-N at different pH values - Table 8.3.7 of ANZECC (2000) - based on average laboratory pH of 7.3 from pH values presented above;

# - # - based on the recalculated trigger value for freshwater, Hickey 2013; and

values above the guidelines are **bolded**.

**Table 3: Surface water results – May 2019**

Sample	NH4-N	HCO3	Sol. Fe	Tot Fe	FCs	E. Coli
Units	mg/L	mg/L	mg/L	mg/L	CFU/100ml	CFU/100ml
LP1	<b>895</b>	5750	3.7	7.9	ND	ND
SWC2	<b>3.0</b>	210	0.13	1.5	---	---
SWC-UP	<0.1	165	0.16	16	---	---
SWC-Down	0.4	195	0.09	0.24	---	---
SWC_DOWN_2	0.3	190	0.05	0.32	---	---
SWP1	0.3	175	0.64	1.0	---	---
SWP2	0.1	545	0.07	0.09	---	---
SWP4	3.2	435	0.09	0.35	---	---
<b>ANZECC 2000</b>	1.88*	---	0.3#	---	---	---

Notes:

--- = not analysed;

ND: not detected

FCs = faecal coliforms;

E. Coli = Escherichia coli;

Guidelines levels from ANZECC (2000) – *Australian and New Zealand guidelines for fresh and marine water quality for the protection of aquatic ecosystems*;

\* = guideline from marine trigger values as total NH4-N at different pH values - Table 8.3.7 of ANZECC (2000) - Table 8.3.7 of ANZECC (2000) - based on average laboratory pH of 7.3 from pH values presented in Table 1;

# = interim indicative working level presented in section 8.3.7 of ANZECC 2000 (based on Canadian derived guidelines); and

Values above the guidelines are **bolded**.

**Table 4: Ratios of principal ions – May 2019**

Bore	Na/Cl	Na/Ca	Mg/Ca	Ca/K	Cl/SO <sub>4</sub>	Cl/HCO <sub>3</sub>	K/TDS	L/N
Units							%	%
BH1c	1.18	4.36	1.46	1.15	76.78	0.48	5.49	60.91
BH2	1.29	1.67	0.70	7.59	4.52	0.56	2.47	14.74
BH3	0.54	0.34	0.25	10.03	3.52	0.70	4.52	60.29
BH4	1.04	0.67	0.31	15.44	1.88	0.55	2.18	8.93
BH13	1.56	0.50	0.40	9.75	0.61	0.26	3.52	12.03
BH14	1.35	0.92	0.48	19.02	3.75	0.44	1.55	5.14
BH16	1.16	1.80	1.55	3.95	1.39	0.86	3.67	12.66
BH20	0.63	0.24	0.31	10.97	0.68	0.47	3.74	17.54
BH20s	1.31	0.33	0.54	2.69	0.39	0.22	10.74	69.32
LP1	1.25	10.30	1.39	0.60	20.65	0.48	4.95	83.50
SWP1	1.13	1.24	0.64	7.37	7.39	0.59	3.04	10.33
SWP2	1.31	3.01	1.02	5.85	2.63	1.04	2.14	6.84
SWP4	1.44	4.92	1.65	5.26	1.67	1.48	1.70	5.70
SWC-UP	0.86	16.77	3.93	3.06	9.27	84.18	1.01	2.82
SWC-DOWN	0.84	19.38	4.51	2.34	9.98	108.57	1.11	3.22
SWC_DOWN_2	0.82	19.54	4.47	2.28	9.90	117.76	1.12	3.28

1. Notes:

% indicates ratios are presented in percentage in that column; and  
L/N = leachate/non-leachate ratio ; [(K + NH<sub>4</sub> + NO<sub>3</sub> + NO<sub>2</sub>)/(Ca + Mg + Na)] x 100.

**Table 5: Summary of gas analysis – May 2019**

Location	GA 5000	ILU
Units	V/V%	ppm
Landfill cap	0.0	Max: 206 / Ave: 15.79
Main weigh bridge, weigh bridge office and landfill office sheds	0.0	Max: 9.8 / Ave: 5.63
Dunmore Resource & Recycling Services	0.0	Max: 3.3 / Ave: 2.47
<b>GUIDELINES</b>	1.25 % v/v / 0.05 % v/v	500 ppm / 0.05 v/v

Notes:

Results and guidelines are expressed in V/V %;

Guidelines are as per the NSW EPA (1996) reporting accumulation value of 1.25 % v/v CH<sub>4</sub>; and surface emission trigger value (500 ppm or 0.05 % v/v); and

Values above the guidelines are **bolded**.

**Table 6: Summary of gas analysis – May 2019**

Analytes	BH4	FD1	RPD(%)
pH	7.1	7.2	1.40
TDS	1100	1130	2.69
Na <sup>+</sup>	145	140	3.51
Ca <sup>++</sup>	190	200	5.13
Mg <sup>++</sup>	36	36	0.00
K <sup>+</sup>	24	25	4.08
NH <sub>4</sub> -N	8.9	9.3	4.40
Cl <sup>-</sup>	215	215	0.00
SO <sub>4</sub> <sup>--</sup>	155	150	3.28
HCO <sub>3</sub> <sup>-</sup>	675	670	0.74
NO <sub>3</sub> <sup>-</sup>	<0.1	<0.1	0.00
PO <sub>4</sub> <sup>---</sup>	<0.1	<0.1	0.00
F <sup>-</sup>	<0.1	<0.1	0.00
BOD	3	<2	NC
Fe.D	0.19	0.22	14.63
Fe.T	5.6	5.6	0.00
Mn.D	0.21	0.23	9.09
TOC	20	20	0.00

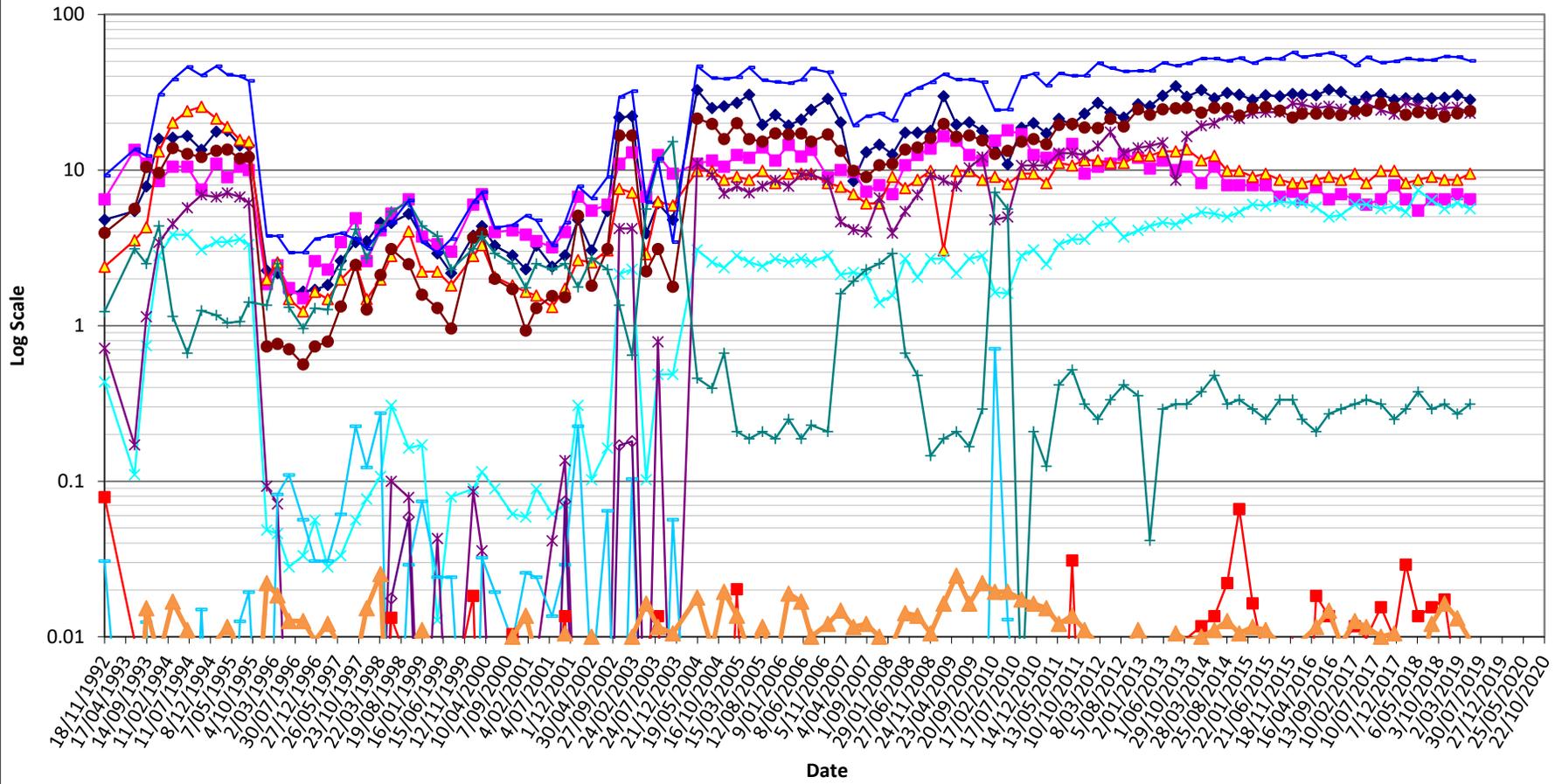
Notes:

Results are expressed in mg/L;  
 NC: not calculated  
 RPD – Relative Percentage Difference  
 Values requiring further investigation are **bolded**.

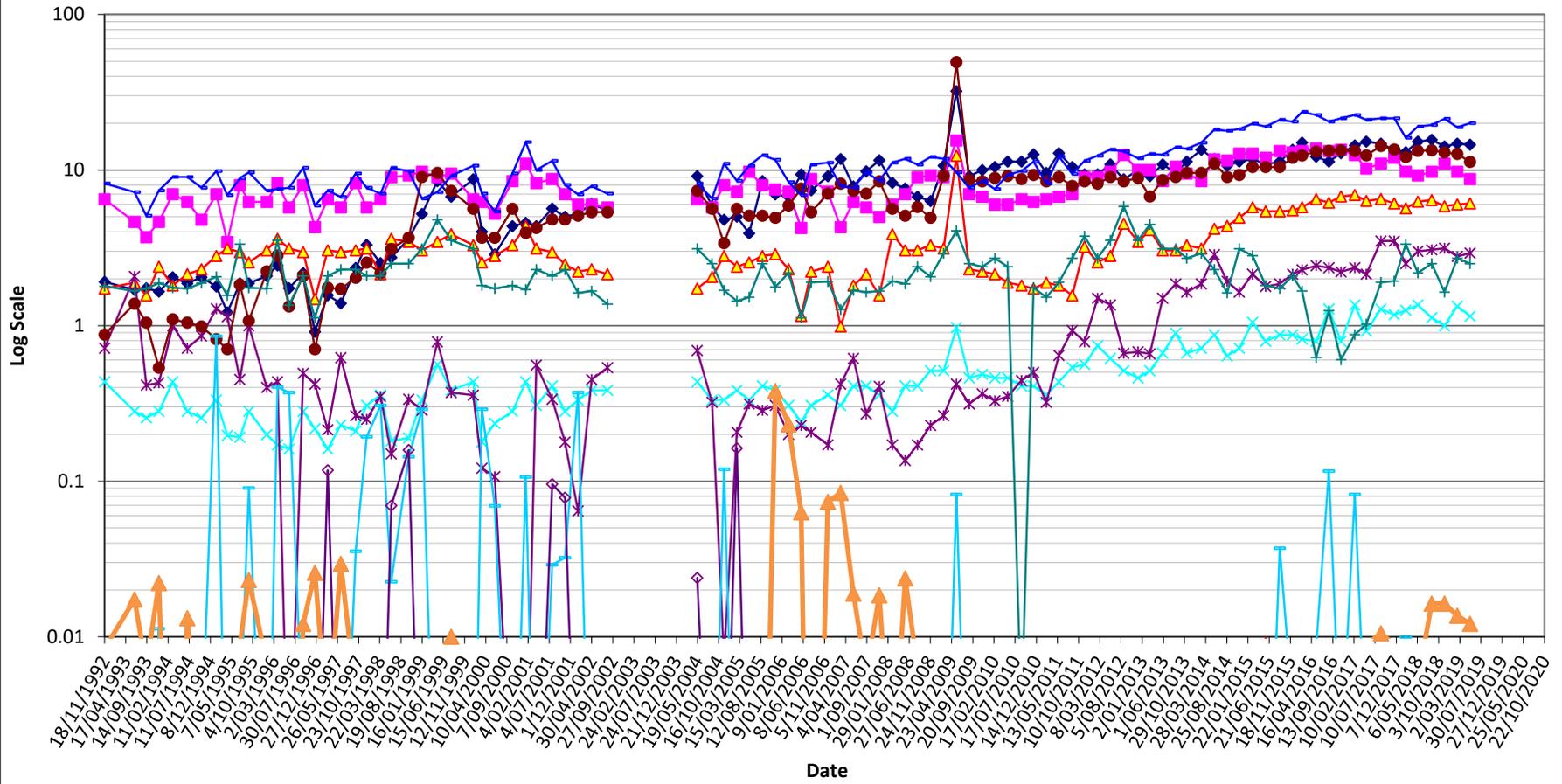
## APPENDIX C: SCHOELLER PLOTS

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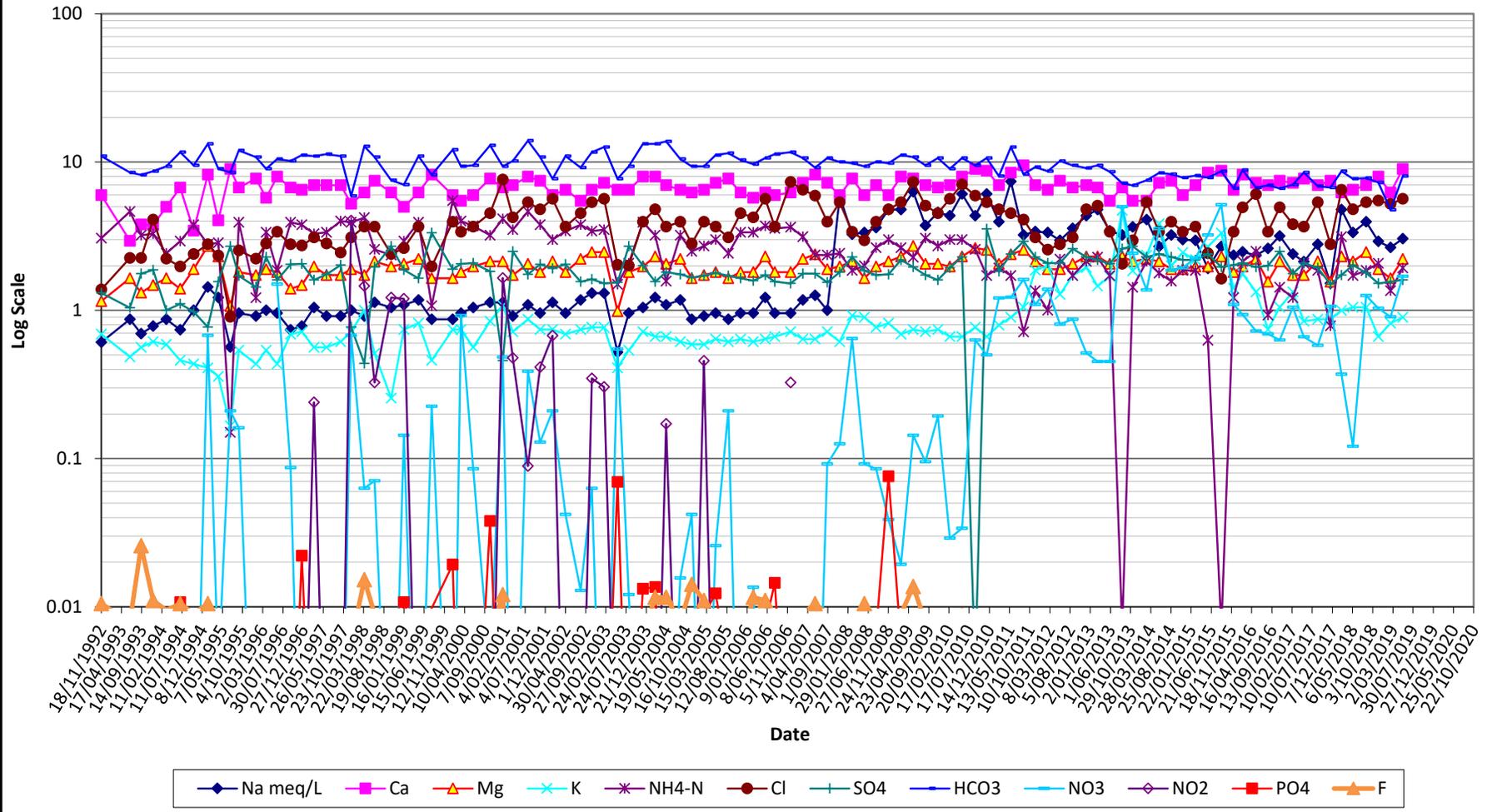
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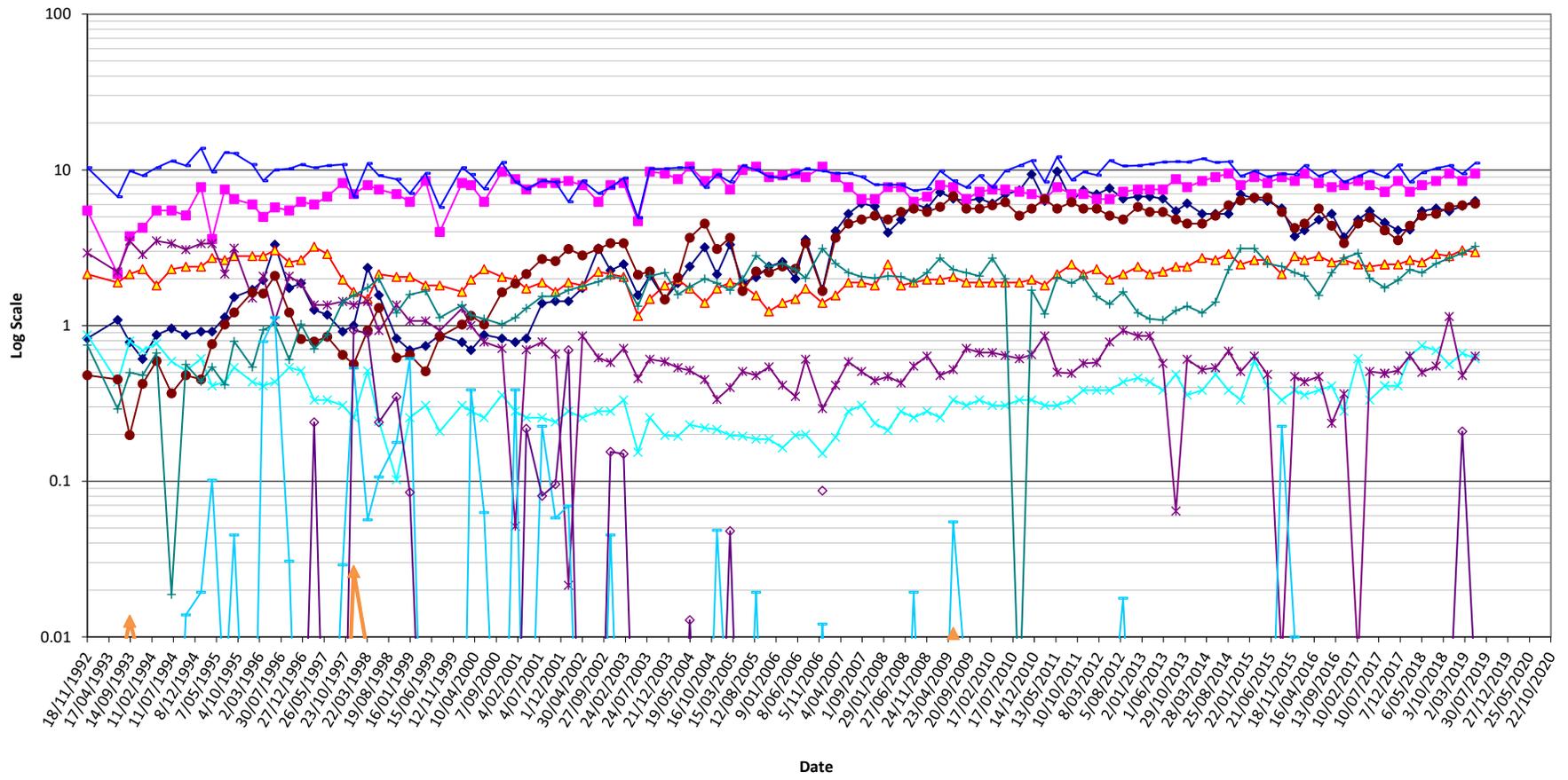
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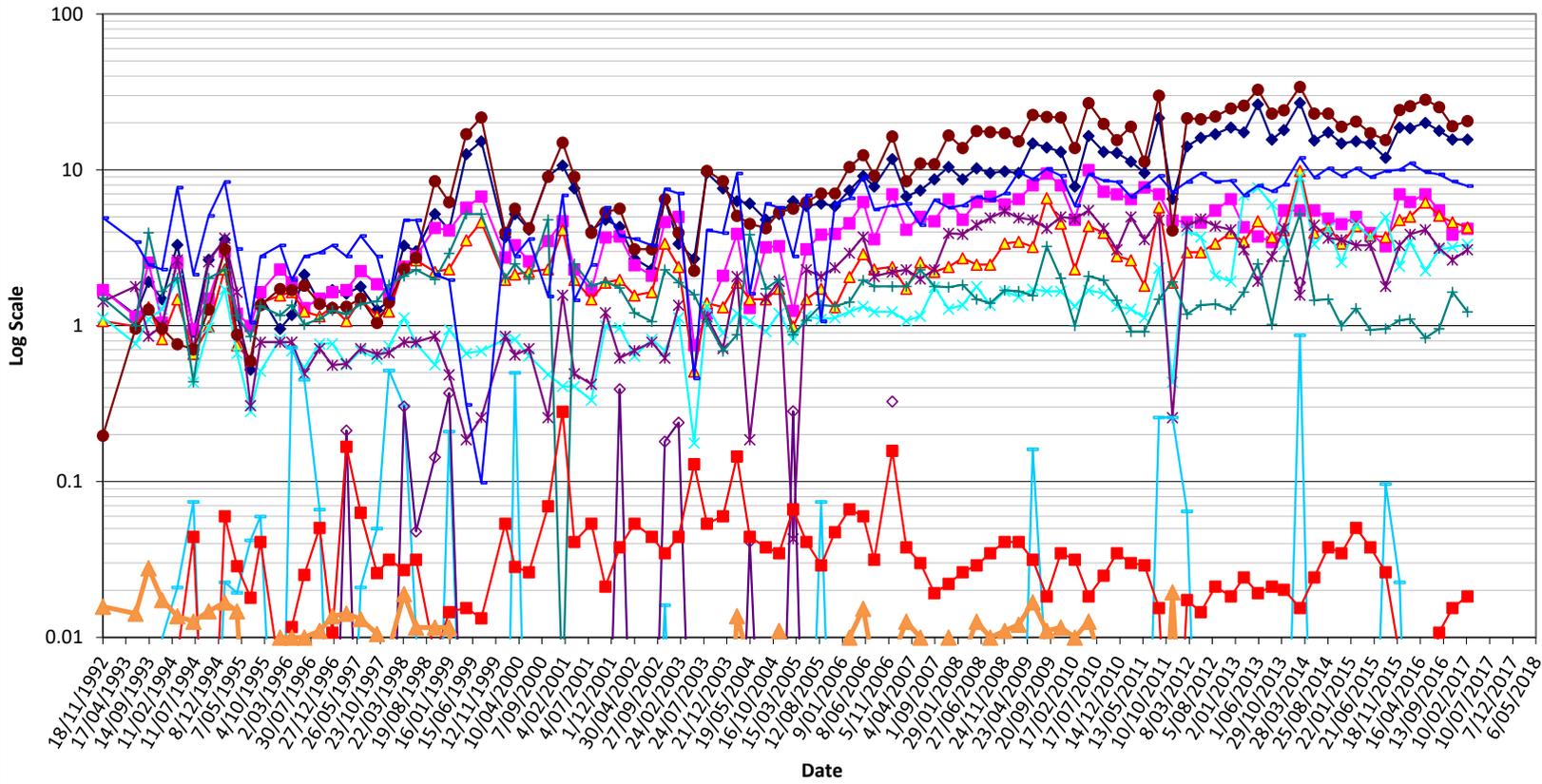
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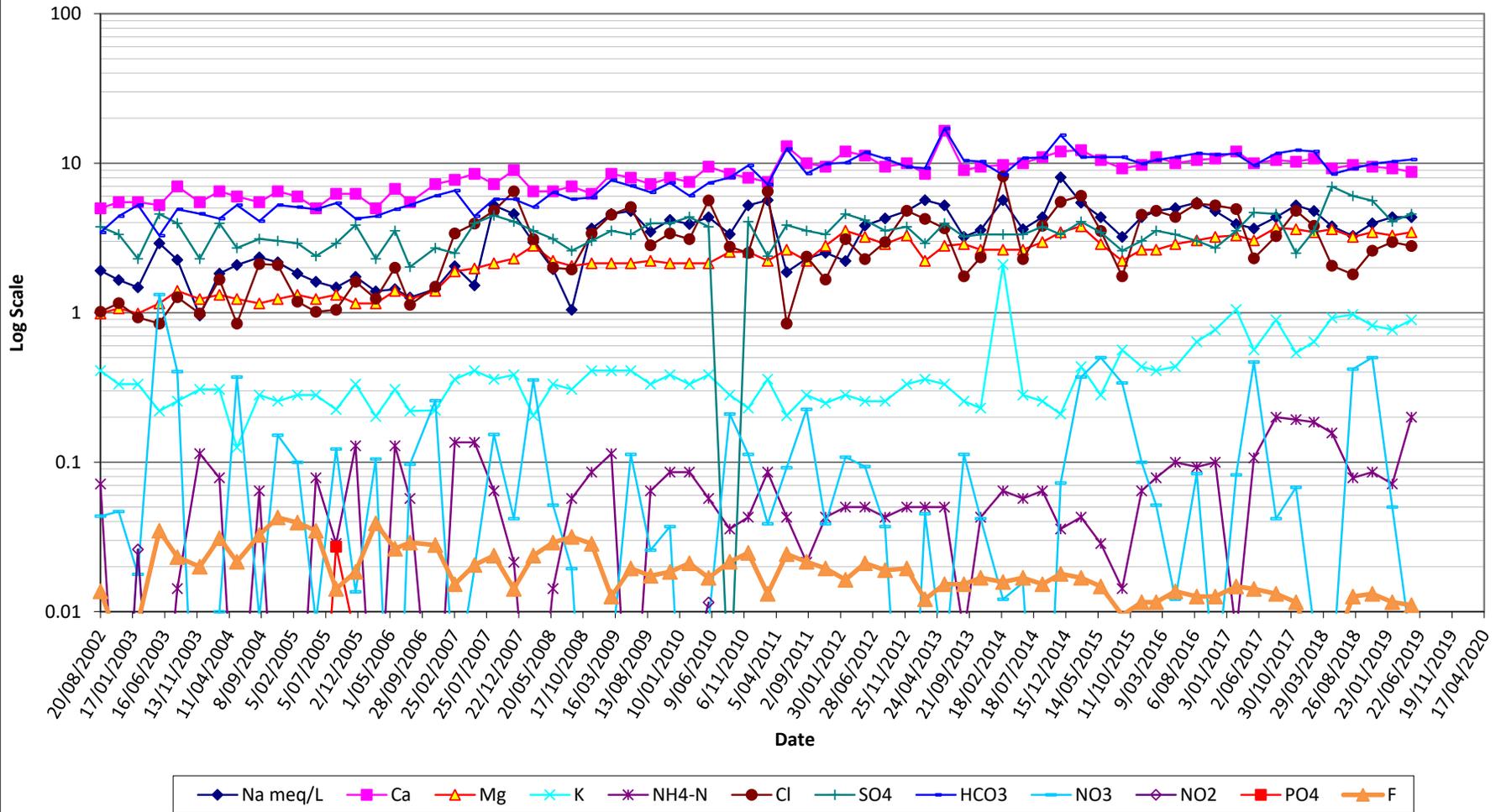
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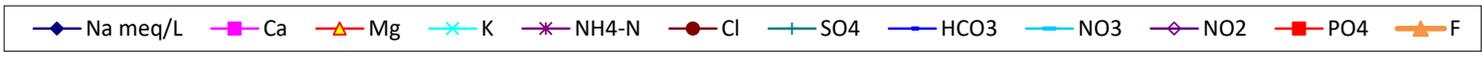
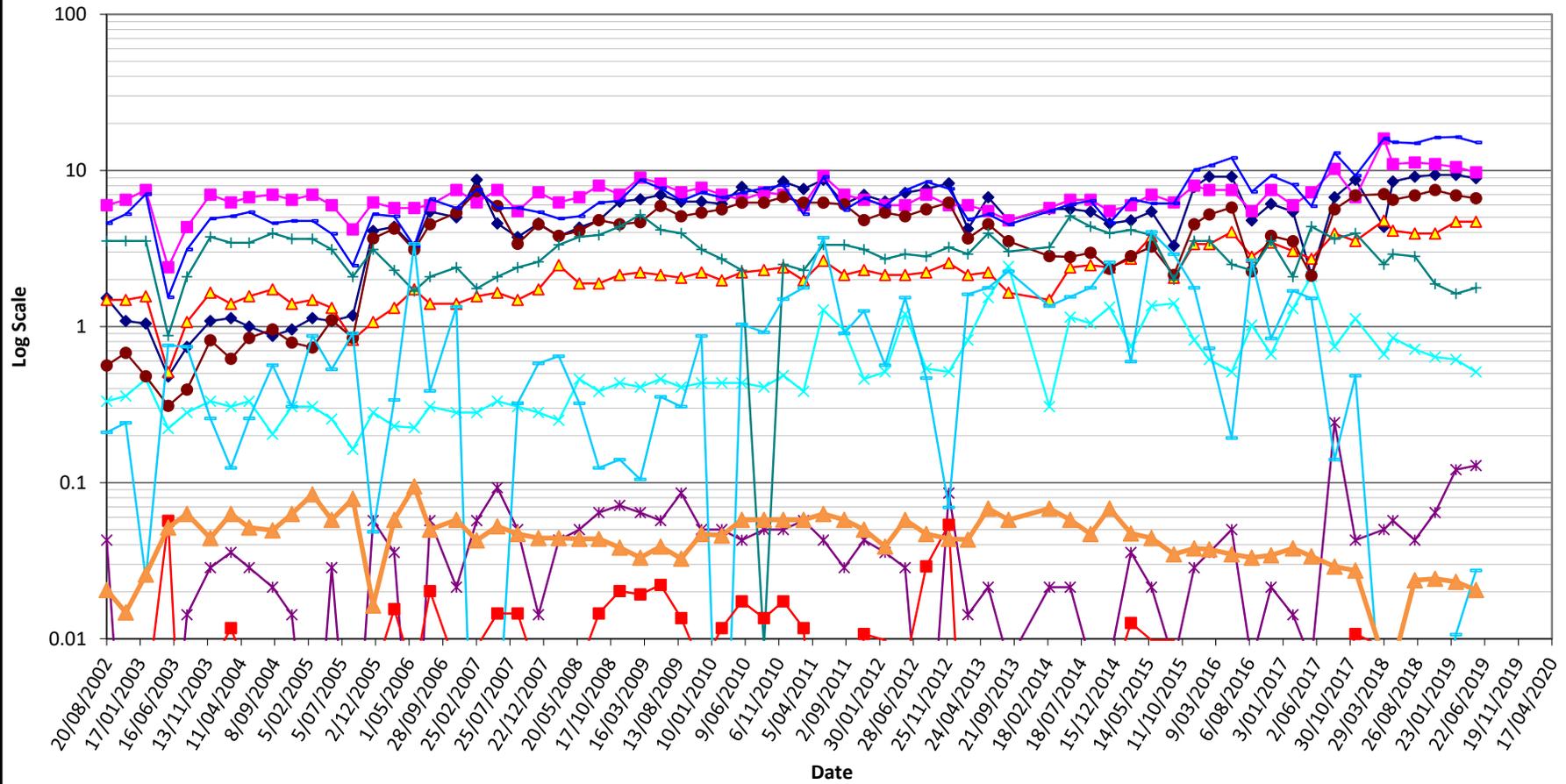
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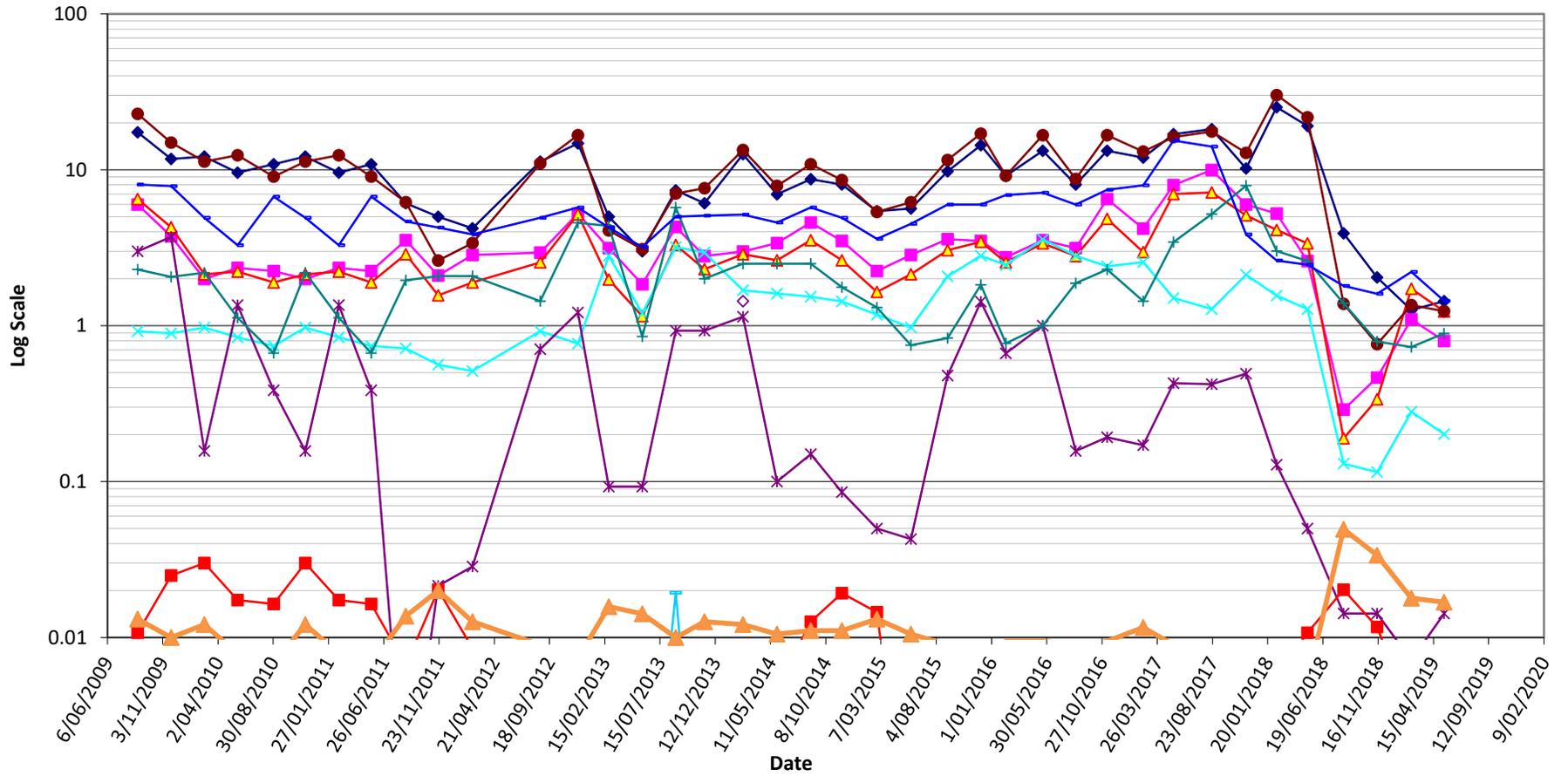
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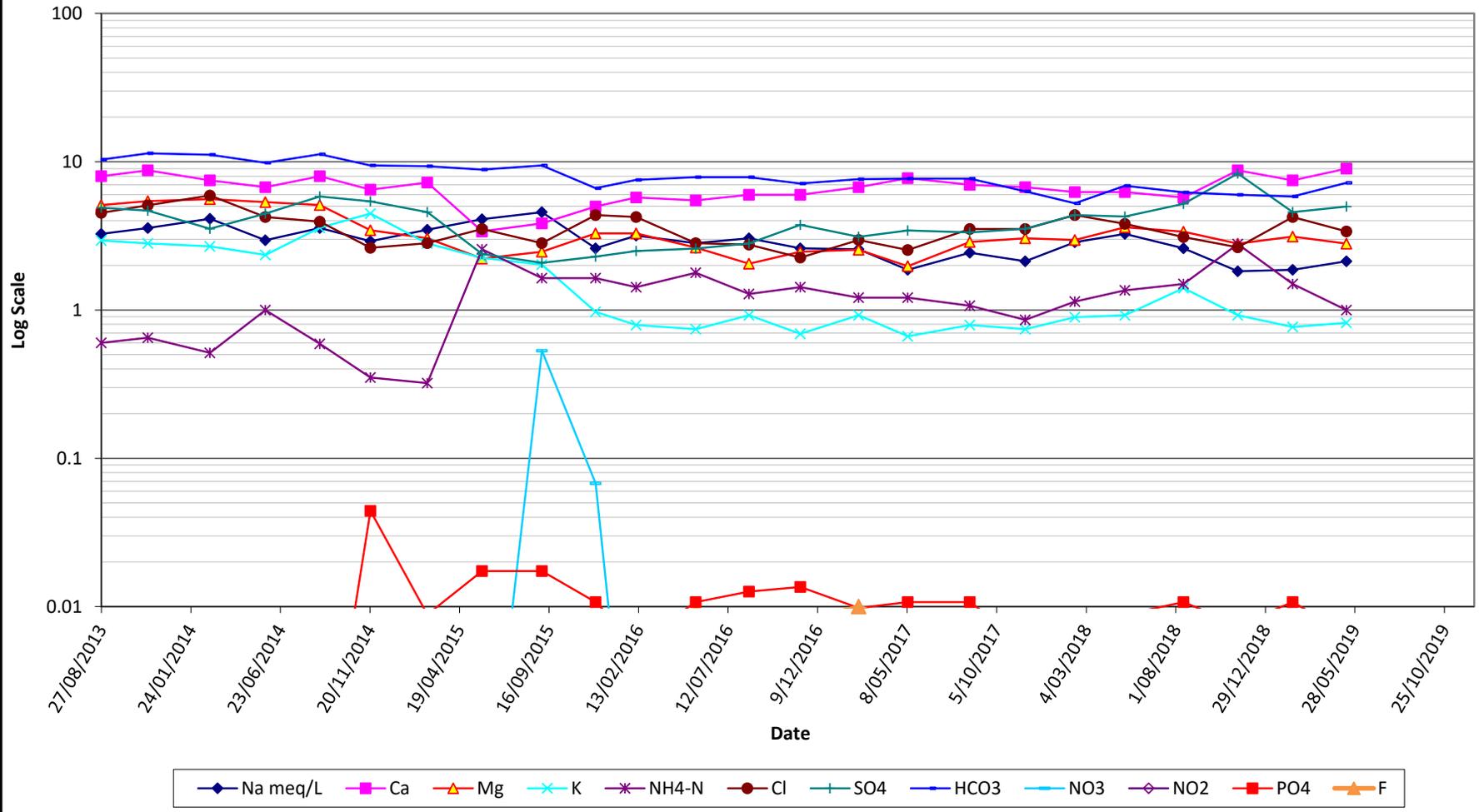
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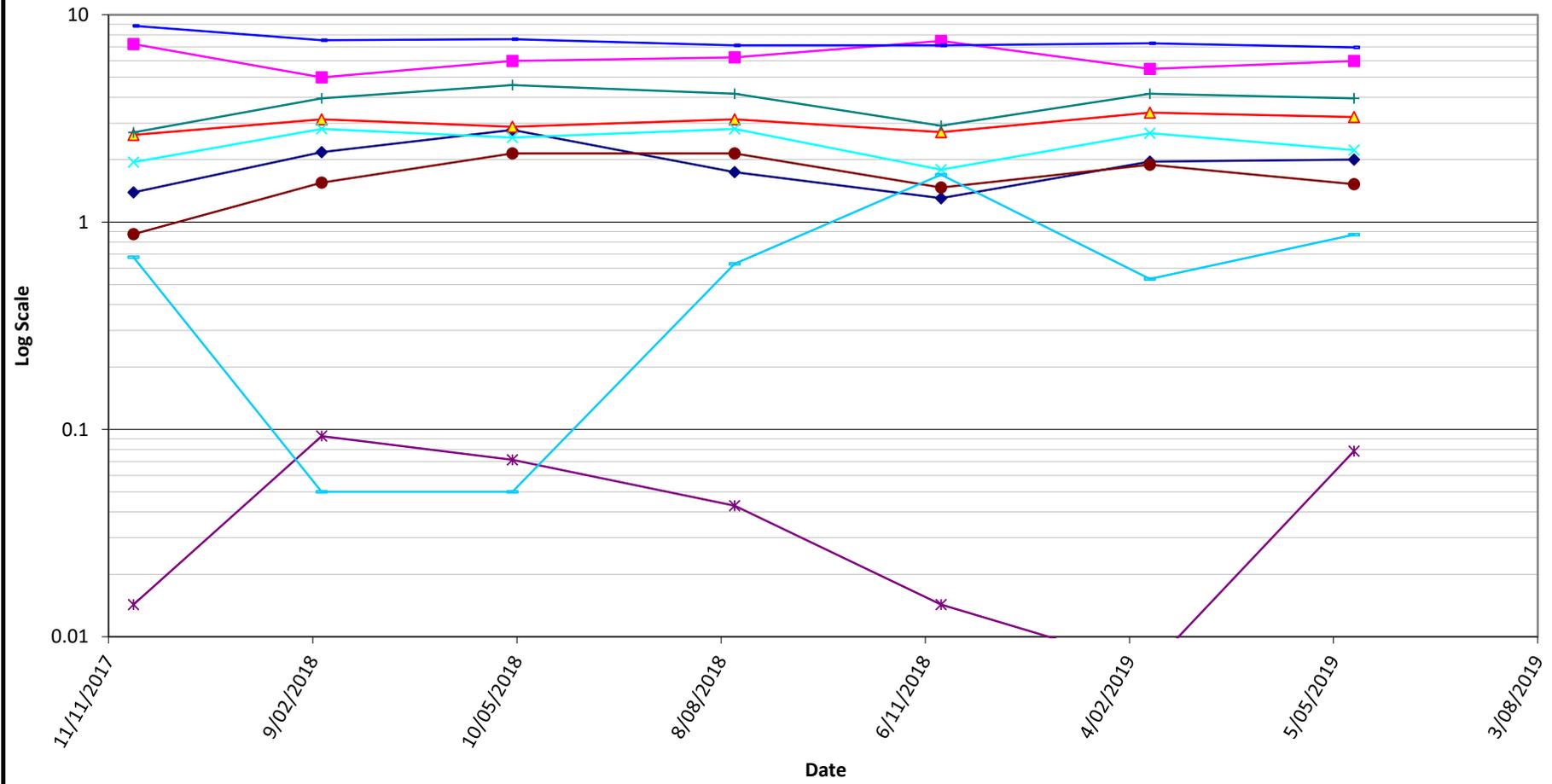
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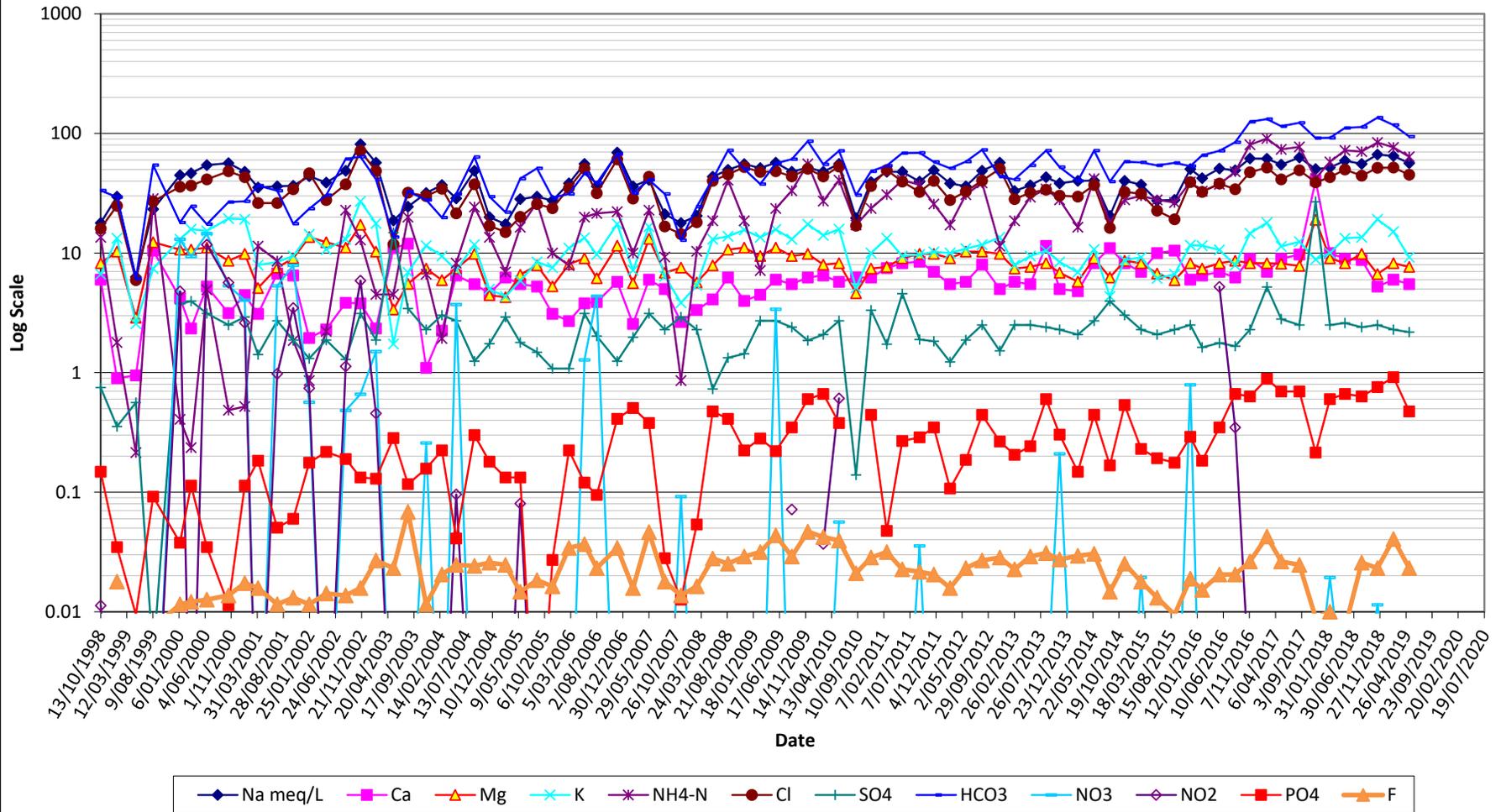
# BH20



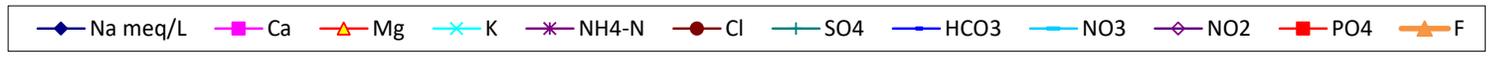
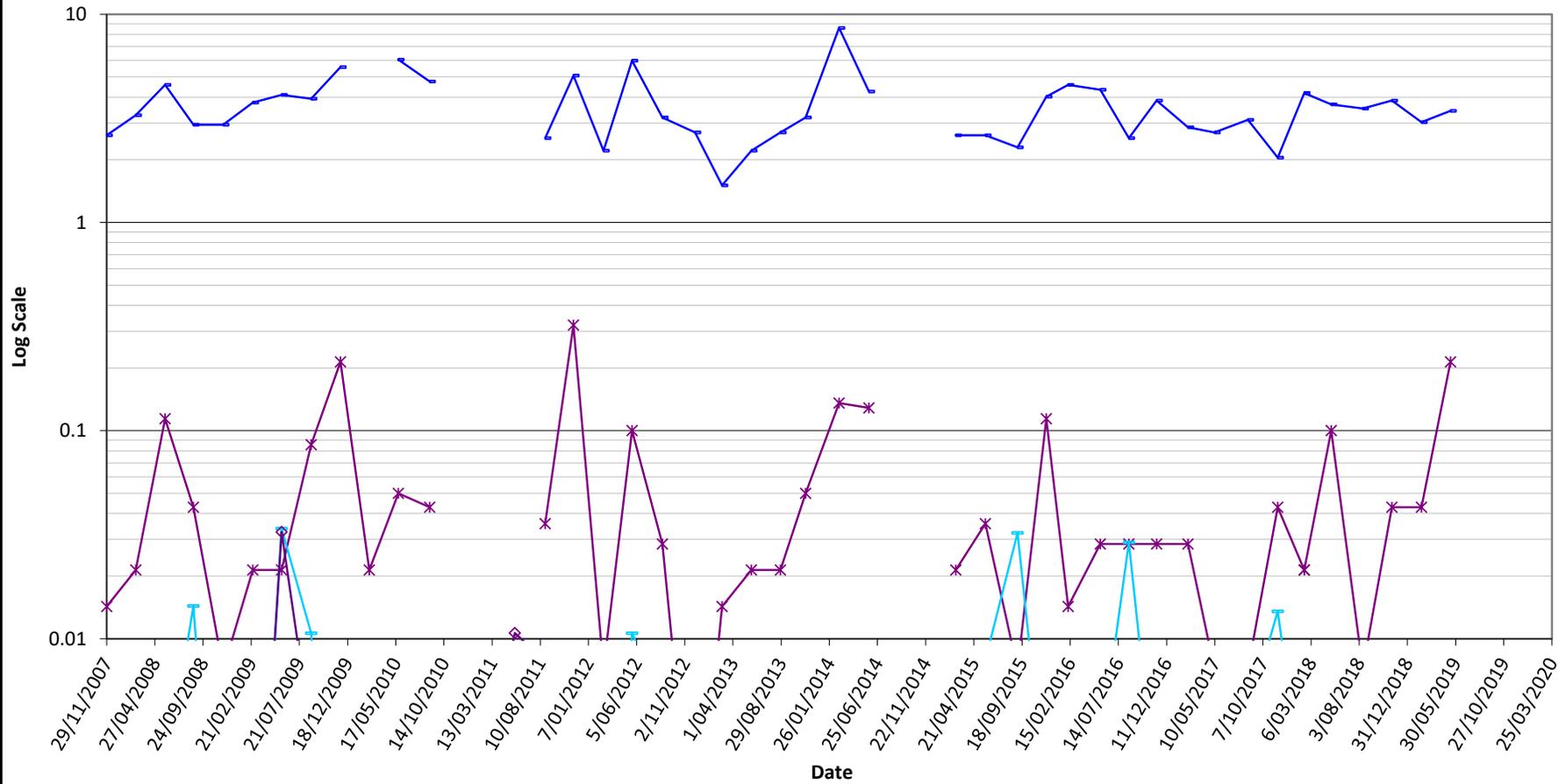
# BH20s



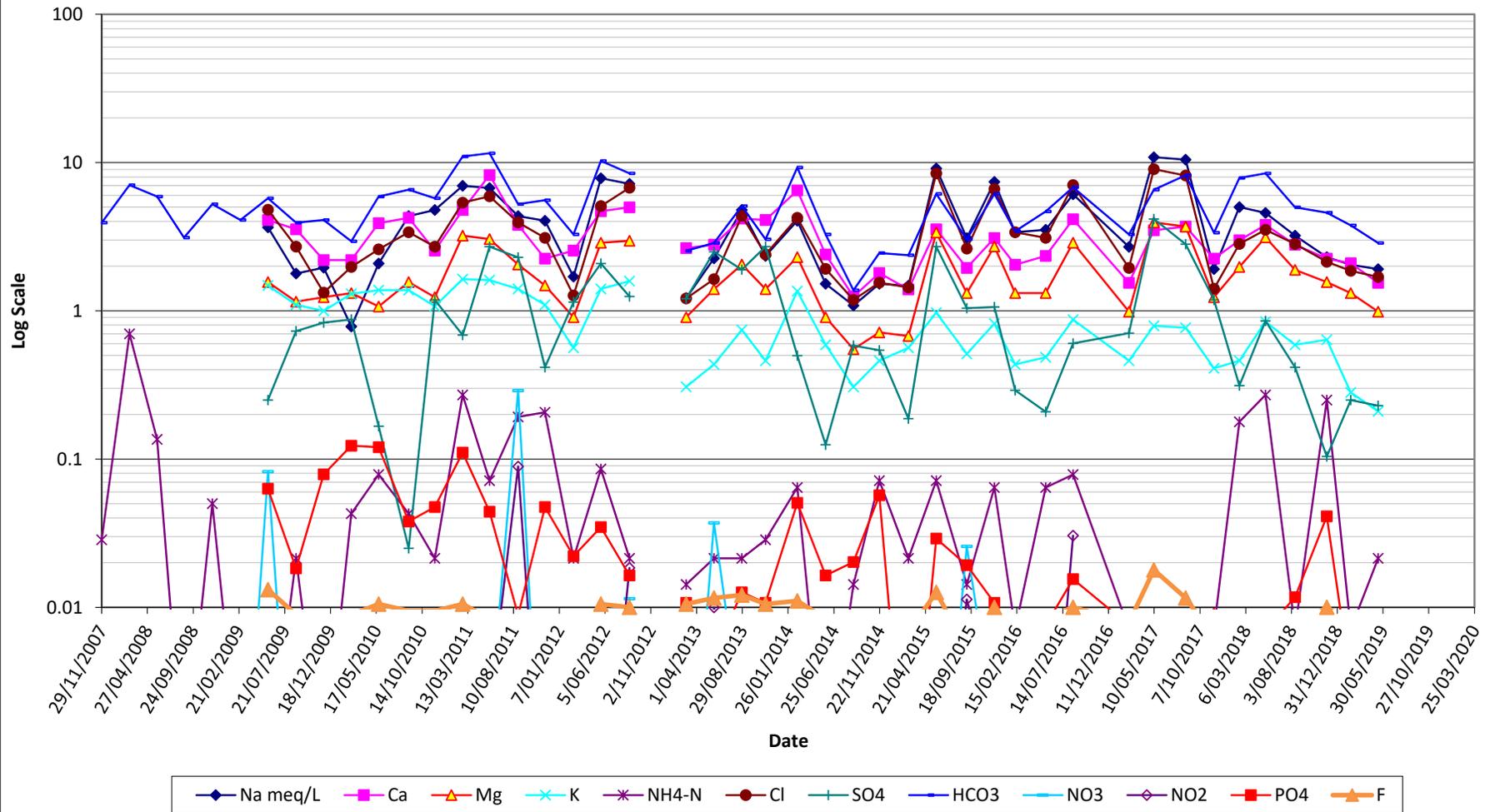
# LP1



# SWC2

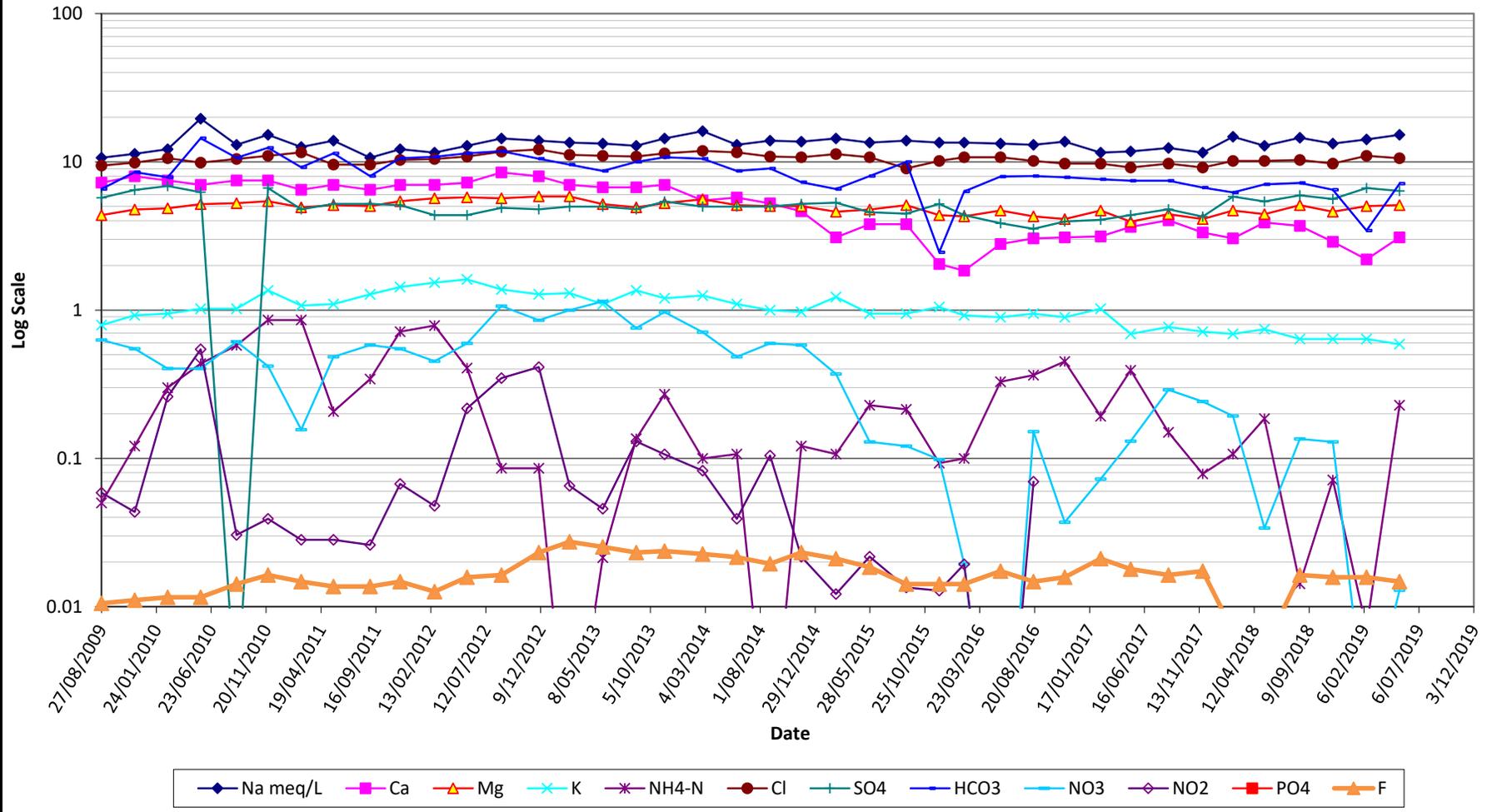


# SWP1

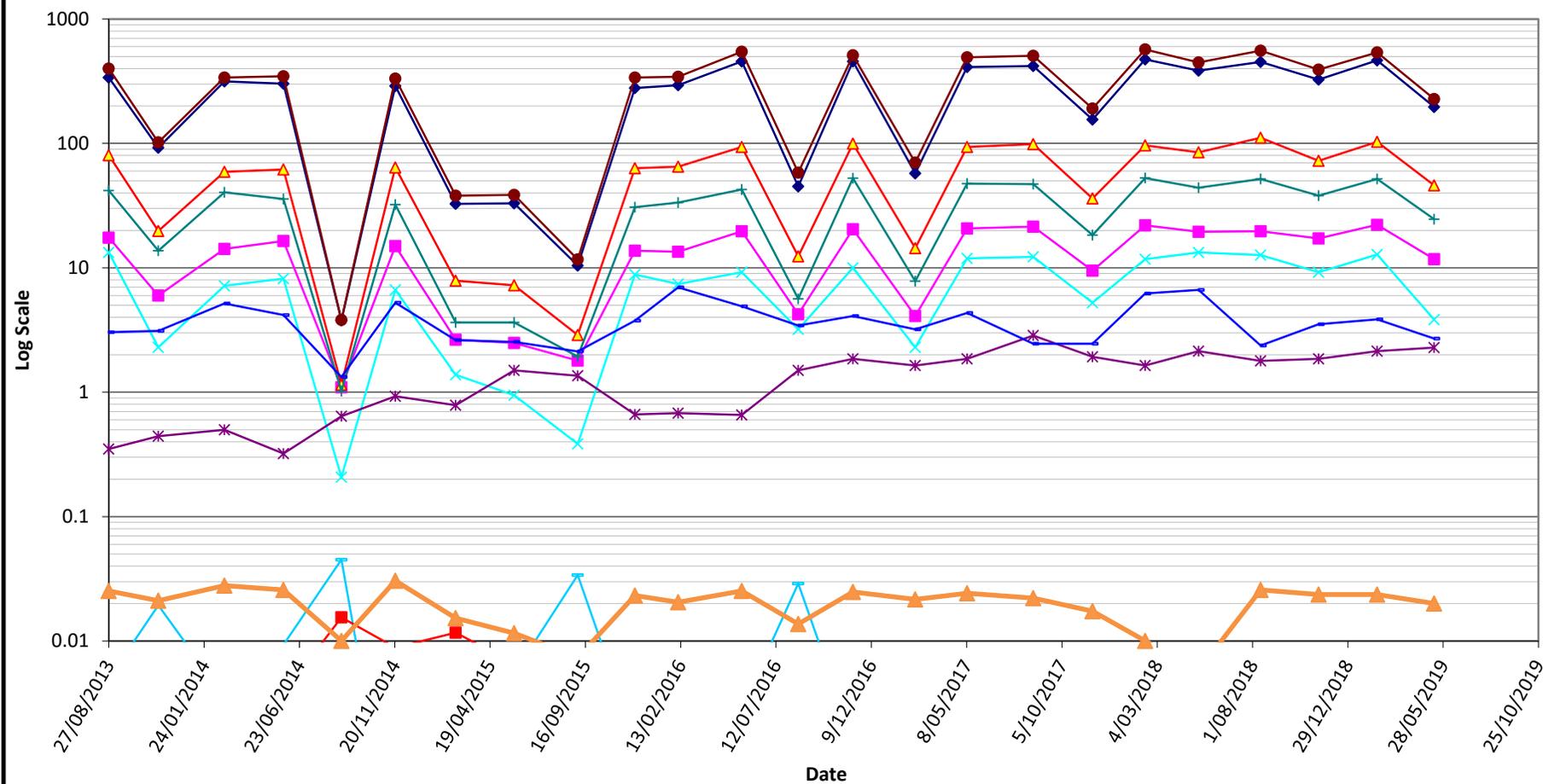




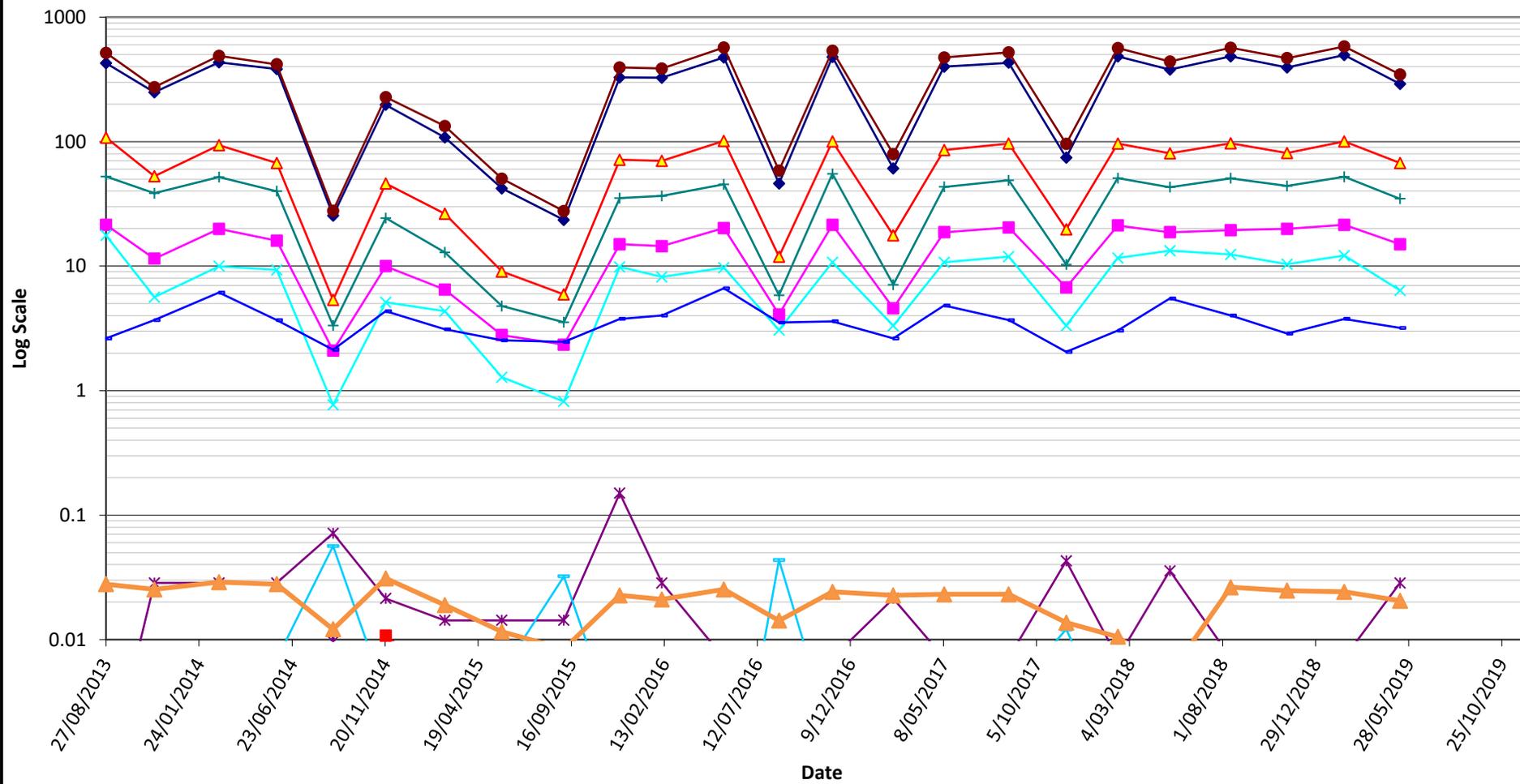
# SWP4



# SWC\_Up



# SWC\_Down



# SWC\_Down\_2

